

**AN ASSESSMENT OF CHANGES IN THE ICHTHYOFAUNAL BYCATCH OF THE TUGELA BANK
PRAWN TRAWLERS IN KWAZULU NATAL**

By

Mbali Mkhize (203513732)

Submitted in partial fulfilment of the academic
Requirements for the degree of
Master of Marine and Coastal Management in the
School of Life and Environmental Sciences
University of KwaZulu-Natal, Howard College and the
Oceanographic Research Institute
Durban

March 2006

PREFACE

The work in this thesis was carried out at the Oceanographic Research Institute in Durban, under the supervision of Dr S.T. Fennessy (Senior Scientist) and Professor R. van der Elst (Director) from February 2003 to February 2005.

This study represents original work by the author and has not been submitted in any other form to another University. Where use was made of the work or research samples of others it has been duly acknowledged in the text.

ACKNOWLEDGEMENTS

This study was funded by Marine and Coastal Management as well as the South African Association for Marine Biological Research and I am very grateful to them for making this possible.

I would also like to personally thank the following people:

Dr Sean Fennessy who has played a very vital role in the completion of this study, as he was the one reading countless drafts trying to get the final product perfect. Thank you for the weekends you put in reading and correcting my work, my writing skills are the better because of your input. I would like to thank Risha Persad, my partner in sorting all the samples from the prawn trawlers, it was a smelly job, but someone had to do it. To Nico and the other observers from Capricorn Fisheries Monitoring who went onboard the trawlers for weeks on end to bring back the samples for me. Many thanks to Nat for helping me with the shark and ray species identification. The University of KwaZulu-Natal for awarding me with a Graduate Assistant Bursary to study at the university.

I would also like to thank the South African Institute of Aquatic Biodiversity (SAIAB) and IZIKO Natural History Museum (Cape Town) for the identification of some of the specimens.

ABSTRACT

Bycatch refers to the portion of the catch that is captured incidentally to the target species and is one of many growing problems facing the world's marine fisheries in recent years. Prawn trawling is one of the world's most lucrative marine fisheries, contributing about 3% to the total annual production of the world's marine fisheries. It is also one the most wasteful because it is associated with large quantities of bycatch. This is because of the high diversity and abundance of other organisms in areas where prawn trawling occurs, and the non-selective nature of the otter trawl used to land prawns. The South African shallow water prawn trawl fishery is typical of penaeid fisheries, and operates on the Tugela Bank off northern KwaZulu-Natal. The bycatch of penaeid prawn trawlers operating on the Tugela Bank was analysed between March 2003-July 2004 as a follow up to an initial study between May 1989-June 1992. The aim of this study was to determine if there has been any change in bycatch species composition over the 13-years and to determine the impacts of trawling on the environment.

Catch composition data were recorded from 168 trawls processed onboard trawlers and on-shore. A total of 122 species was identified with teleosts contributing more than 60% to relative abundance by number. Comparison between the 1989-1992 and 2003-2004 data sets showed that although the species compositions were similar, there was an increase in the relative abundance of pelagic species. It is not clear if the change is due to trawling impacts or due to fluctuations in recruitment of these taxa.

The annual discarded bycatch was estimated at 88 tonnes in 2003 and the prawn catch was estimated at 11 tonnes, both substantially lower than in 1989-1992. There was no difference in discard rate between trawls from shallow and medium depths. However, there were significant differences in discard rates between large and medium gears, and warm (December-April) and cool (May-November) seasons. Kolmogorov-Smirnov tests revealed that there were significant differences between the 1990s and 2000s length frequency distributions of three sciaenid species (*Otolithes ruber*, *Johnius dorsalis* and *J. fuscolineatus*), which all showed an increase in fish size for the 2003-2004 data. This assessment, however, is confounded by a gear effect as mesh size differs between the two sample periods.

Based on the results of the study, management recommendations were made to improve the management of the Tugela Bank prawn trawl fishery by reducing bycatch. It was recommended that the observer programme continue for further investigation into the changes in species dynamics. Further research on BRDs (Bycatch Reducing Devices) and gear types that reduce bycatch is also recommended.

TABLE OF CONTENTS

PREFACE	II
ACKNOWLEDGEMENTS	III
ABSTRACT	IV
TABLE OF CONTENTS	V
1. INTRODUCTION	1
2. STUDY AREA	3
3. DESCRIPTION OF THE TUGELA BANK PRAWN TRAWL FISHERY	5
4. MATERIALS AND METHODS	6
5. RESULTS	9
6. DISCUSSION	19
7. MANAGEMENT OF THE FISHERY AND RECOMMENDATIONS	25
REFERENCES	28
APPENDIX	35

1. INTRODUCTION

Commercial prawn trawling occurs in the majority of the world's oceans, and provides a total catch of up to 3 million tonnes per year which accounts for about 3.2% of the total production from the world's marine fisheries (FAO, 2002). Many marine fisheries, including prawn trawling are faced with the problem of bycatch. Bycatch refers to the portion of the catch that is caught incidental to the target species (Saila, 1983), and it has played a significant role in the decline of many of the world's fish stocks because high catches of juveniles of many bycatch species. Bycatch makes up approximately 10% of the total annual global landings (Japp, 1997). Discarded bycatch makes up 90% of what is caught in prawn trawls world wide, with only 10% of the catch being retained (Smith, 2000). The high percentage of discarded bycatch in prawn trawls is a result of the poor selectivity of the otter trawl, which is the gear most commonly used to catch prawns (Vendeville, 1990). The issue of bycatch in commercial fisheries has presented serious problems and has proved to be one of the most important issues facing the fishing industry to date (Kennelly *et al.*, 1998).

A major concern of bycatch is the catching of species that are targeted in other fisheries (Vaughan *et al.*, 1991; Kennelly *et al.*, 1993), the consequence of which is the reduction of the recruitment, biomass and yield of these stocks. This is largely because of the high mortality of their juveniles and subadults which are caught as bycatch (Andrew and Pepperell, 1992; Broadhurst, 2000). This threatens the maintenance of biodiversity and long term sustainability of marine ecosystems, and contributes to ecosystem degradation (FAO, 1995).

Penaeid (prawns with a juvenile estuarine stage in their development) fisheries are one of the world's most lucrative because of the high value of the catch, but they are also one of the most wasteful because of the large quantities of bycatch associated with them (Andrew and Pepperell, 1992). In most of these fisheries, the species that are of commercial value are retained, and the so-called 'trash' component of the bycatch is discarded because of its poor marketability (Andrew and Pepperell, 1992; Fennessy, 1994a).

Most of the world's penaeid prawn fisheries are situated on the continental shelf in tropical or sub-tropical regions where there is a high abundance and diversity of other organisms (van der Elst, 1988; Fennessy, 1994b). The South African prawn trawl fishery occurs in sub-tropical KwaZulu-Natal and consists of shallow and deep water fisheries (Fennessy and Groeneveld, 1997). The shallow water fishery, which operates at depths of 20-50 m, is situated on the Tugela Bank trawling grounds off the KwaZulu-Natal north coast and the deep water fishery operates at 100-600 m. There are limited possibilities for penaeid prawn trawling in South Africa because of the narrow width of the continental shelf along most of the South African coastline, and the Tugela Bank grounds are one of

the few suitable areas because the shelf is at it's widest here (Figure 1; Fennessy *et al.*, 1994; Fennessy, 1995). The Tugela Bank coincides with an area of mud deposition from discharges of many rivers in the area, which makes it ideal for penaeid prawns (McCormick *et al.*, 1992). The Tugela Bank supports a seasonal fishery which catches white prawn, *Fenneropenaeus indicus*, but also catches a large amount of teleosts, elasmobranchs, gastropods, cephalopods and other crustaceans as bycatch (Fennessy, 1994b; Fennessy *et al.*, 1994; Fennessy and Groeneveld, 1997).

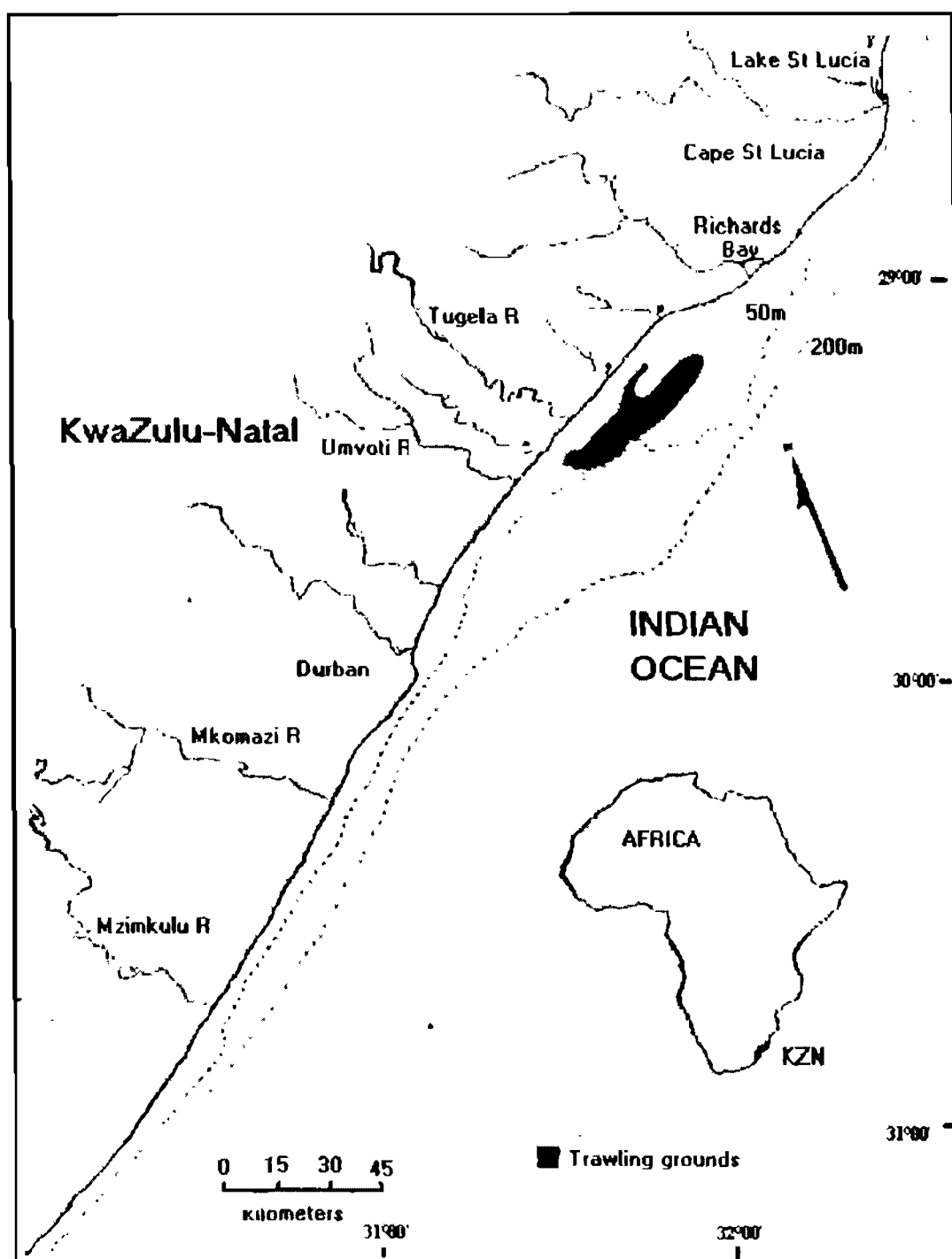
Although the scale of this fishery is small compared to other larger commercial fisheries in South Africa, the bycatch and discard proportions are high in comparison with other South African fisheries (Japp, 1997). There are other important fisheries operating in the Tugela area, such as the skiboat linefishery and the shore linefishery which are either commercial or recreational in nature (Mann *et al.* 1997). Participants in these sectors believe that the trawl fishery is responsible for the decline in species targeted by the other fisheries operating in the area (Fennessy, 1993, Mann *et al.*, 1997). The aim of this study is to investigate the species composition and to quantity of the bycatch from prawn trawlers operating on the Tugela Bank. An initial survey from 1989 to 1992 provided a baseline description of the bycatch (Fennessy *et al.*, 1994; Fennessy, 1993), and the intention of the current study was to repeat the survey after a 13-year period, to attempt to assess the impact that trawlers may have had on the region. This study also aims to provide management recommendations to reduce bycatch.

2. STUDY AREA: DESCRIPTION OF THE TUGELA BANK

The trawl grounds are situated two to four nautical miles offshore from the Matikulu River mouth (29°06'36''S, 31°36'48''E) on the KZN north coast, about 100 km from Durban (Figure 1). The grounds cover an area of approximately 500 km², with trawl depths ranging from 20 to 50 m (Fennessy, 1995). The continental shelf is at its widest here, with the shelf break occurring 50 km offshore due to the gentle Natal Bight found between Durban and St Lucia (Lutjeharms *et al.*, 1989). The hydrodynamics of this area are dominated by wave action, rather than the Agulhas Current, which is forced further offshore by the Bight (Flemming and Hay, 1988). The south flowing Agulhas Current is a dominating dynamic feature that affects the flow along the entire KZN coastal region and it flows in a south-southeasterly direction (Schumann, 1982). There is also a strong counter-current inshore of the Agulhas Current, which flows in a north-easterly direction (Lutjeharms and Connell, 1989). This counter-current is thought to be driven by a vortex shedding from the Natal Bight, known as the Natal Pulse. These pulses originate at the Bight and grow laterally, moving downstream with average diameters of 170 km, whereas the Agulhas Current seldom exceeds 15 km in width (Lutjeharms and Connell, 1989).

The KwaZulu-Natal shelf water temperatures are generally well mixed because of the high-energy coastline (Schumann, 1982). Monthly mean sea surface temperatures measured close to the Tugela Bank ranged from 20 to 24.4°C (Fennessy, 1993). The water temperature measured at 10 m depth on the Bank was 17.75°C which was of 3°C lower than ambient shelf waters, and 8°C lower than the Agulhas Current surface waters (Lutjeharms *et al.*, 1989). The occurrence of cooler waters on the Bank is assumed to be a result of upwelling activity (Lutjeharms *et al.*, 1989), and re-circulation in the area (Schumann, 1982). In summer temperatures in the centre of the Bight are 18°C with surface temperature exceeding 22°C (Lutjeharms *et al.*, 1989). North-easterly and southerly winds predominate and they can attain speeds of up to 35 and 60 knots respectively (Hunter, 1988); these contribute to the high energies of the KwaZulu-Natal coast.

Figure 1: Location of the Tugela Bank prawn trawling grounds (After Fennessy, 1993). R=River.



3. DESCRIPTION OF THE TUGELA PRAWN FISHERY

Trawling in this fishery started around 1921, during the Fisheries and Marine Biological Survey on the S.S. Pickle, which was commissioned by the Fisheries Survey Committee of the Department of Mines and Industries (Fennessy, 1993). Trawling then occurred sporadically until 1976 when two vessels began more regular trawling (Fennessy, 1993). However, recording of catch and effort data only started in 1985 using a standard logbook system instituted by the Department of Sea Fisheries (Fennessy and Groeneveld, 1997). Regular statistics were only collected from 1988, when at the time, the fishery was comprised of 12 companies and 21 vessels (Fennessy and Groeneveld, 1997). By 1995, the number of trawlers in operation had dropped to only 7 vessels owned by 4 companies (Fennessy and Groeneveld, 1997), while only 3 trawlers were operating in 2004.

Trawling on the Tugela Bank is seasonal, occurring mainly between March and September (Fennessy, 1993; Fennessy *et al.*, 1994), and is closed from November to February, to reduce bycatch of squaretail kob, *Argyrosomus thorpei* (DEAT, 2002). The target species for this fishery is the white prawn, *F. indicus*, but there are catches of brown prawns, *Metapenaeus monoceros*, and tiger prawns, *Penaeus monodon* (Fennessy, 1993; Fennessy *et al.*, 1994). Although Total Allowable Effort (TAE) is controlled by a permit system, there have been no stock assessments done for any of the targeted species for this fishery, but it is believed to be fully exploited (DEAT, 2002).

The vessels used are steel boats with lengths ranging from 24 to 40 m, and generating between 500 and 1000 hp (Fennessy and Groeneveld, 1997). The gear used is the otter trawl, ranging from single nets to boom-operated twin or triple nets, with footrope lengths of 25 to about 70 m (Fennessy and Groeneveld, 1997). The stretched mesh size is 55 mm in both the wings and the codend (Fennessy and Groeneveld, 1997). Trawling occurs on a 24 h basis with trawl speeds of two to three knots and drag duration averaging 4-5 hours. Trawlers carry crews of between 12 and 20, (Fennessy and Groeneveld, 1997) and the duration of trawl trips can be up to one month.

At the end of each trawl the catch is sorted by the crew and the prawn component is graded by size, packed in cartons and blast frozen as soon as possible (Fennessy, 1993). A large portion of the retained catch is prawns but the largest portion of the catch is bycatch, which, if commercially viable, is packed for marketing. This includes crabs, *Portunus sanguinolentus*, cephalopods, *Uroteuthis duvauceli*, *Sepiella cyanea* and several marketable fish species (Fennessy, 1993). The rest of the bycatch, which has a low market value is discarded.

4. MATERIALS AND METHODS

4.1 Sampling procedure

An initial survey of the bycatch species composition of the prawn trawlers operating on the Tugela Bank was done between May 1989-June 1992 (Fennessy, 1993) and the current study between March 2003 and July 2004 was a follow up to this initial survey.

An observer from Marine and Coastal Management (MCM), a directorate of the South African Department of Environmental Affairs and Tourism, was placed onboard the trawlers to collect discarded bycatch samples. There were a maximum of 120 observer days per year allocated for both shallow and deep water trawling. The trawler trips were of 3-4 weeks duration and trawlers sometimes switched from shallow to deep water trawling during a trip depending on the amount of target species available, so there was no control over the regularity of sampling. Four drags were sampled each day, three of which were processed onboard the trawler by the observer and one sample was frozen for on-shore processing. After the crew had removed the target species and the retained bycatch, the observer removed larger elasmobranchs from the bycatch and collected a ± 20 kg sample of the discarded catch. A species identification catalogue was provided to assist with onboard species identification, and the species that were not included in the catalogue were labelled with a name and trawl details for later identification. Both the observer and the skipper recorded depth, position, time of day and duration of each sampled trawl in drag books. The quantities of retained prawn, retained bycatch and discarded bycatch were also recorded by the observer.

The frozen samples of discarded bycatch were taken to the Oceanographic Research Institute (ORI), in Durban for processing and analysis on-shore. Sorting of the samples included a breakdown of the sample by weight and number of all taxa, and recording the total lengths (TL) of at least fifty individuals of the three most commonly recorded teleost species. A third of the crustacean component was sub-sampled when it was too large to count and weigh each individual. Afterwards necessary corrections were made to estimate actual weight and number of crustaceans in that particular sample. Smith and Heemstra (1986) was used for teleosts species identification and Barnard (1950) for most of the crustaceans. Unidentified specimens were sent to the South African Institute of Aquatic Biodiversity (SAIAB) in Grahamstown and IZIKO Natural History Museum in Cape Town for positive identification. The data were captured onto Microsoft Excel spreadsheets, from which tables of the catch composition were produced illustrating the occurrence and relative contributions by weight and number of the teleosts, invertebrates and elasmobranchs to the total bycatch.

4.2 Analysis of data

Much of the data is descriptive, and is presented in the form of comparative tables. A more rigorous approach, such as multivariate analysis, was considered and rejected, since there was little control over sampling, which was ad hoc (see results).

4.2.1 Total discarded bycatch and retained bycatch

The total annual discarded bycatch for 2003 was estimated by scaling up discarded catch rates (kg/hr) using total fleet effort. The fleet effort, weights of the target species and retained bycatch were obtained from information recorded in drag books by skippers, which were submitted to Marine and Coastal Management. Only 2003 information was available at the time of the analysis.

4.2.2 Differences in discard rates

The effects of gear size, season, depth and day or night on discard rates (kg discard.trawl hour⁻¹) were tested using t-tests. These differences were tested for all the trawls for which the observer had recorded discard rates in the 2003-2004 study period.

4.2.3 Length frequencies

The total lengths of the three most commonly recorded and consistently discarded bycatch species (*O. ruber*, *J. fuscolineatus* and *J. dorsalis*) were measured using a measuring board. A maximum of 50 randomly selected individuals of each species were measured per sample. The data were captured onto Excel worksheets and length frequency distribution histograms were produced. The Kolmogorov-Smirnov two sample test was applied to the data sets from 1989-1992 and 2003-2004 to test for any differences.

1. INTRODUCTION

Commercial prawn trawling occurs in the majority of the world's oceans, and provides a total catch of up to 3 million tonnes per year which accounts for about 3.2% of the total production from the world's marine fisheries (FAO, 2002). Many marine fisheries, including prawn trawling are faced with the problem of bycatch. Bycatch refers to the portion of the catch that is caught incidental to the target species (Saila, 1983), and it has played a significant role in the decline of many of the world's fish stocks because high catches of juveniles of many bycatch species. Bycatch makes up approximately 10% of the total annual global landings (Japp, 1997). Discarded bycatch makes up 90% of what is caught in prawn trawls world wide, with only 10% of the catch being retained (Smith, 2000). The high percentage of discarded bycatch in prawn trawls is a result of the poor selectivity of the otter trawl, which is the gear most commonly used to catch prawns (Vendeville, 1990). The issue of bycatch in commercial fisheries has presented serious problems and has proved to be one of the most important issues facing the fishing industry to date (Kennelly *et al.*, 1998).

A major concern of bycatch is the catching of species that are targeted in other fisheries (Vaughan *et al.*, 1991; Kennelly *et al.*, 1993), the consequence of which is the reduction of the recruitment, biomass and yield of these stocks. This is largely because of the high mortality of their juveniles and subadults which are caught as bycatch (Andrew and Pepperell, 1992; Broadhurst, 2000). This threatens the maintenance of biodiversity and long term sustainability of marine ecosystems, and contributes to ecosystem degradation (FAO, 1995).

Penaeid (prawns with a juvenile estuarine stage in their development) fisheries are one of the world's most lucrative because of the high value of the catch, but they are also one of the most wasteful because of the large quantities of bycatch associated with them (Andrew and Pepperell, 1992). In most of these fisheries, the species that are of commercial value are retained, and the so-called 'trash' component of the bycatch is discarded because of its poor marketability (Andrew and Pepperell, 1992; Fennessy, 1994a).

Most of the world's penaeid prawn fisheries are situated on the continental shelf in tropical or sub-tropical regions where there is a high abundance and diversity of other organisms (van der Elst, 1988; Fennessy, 1994b). The South African prawn trawl fishery occurs in sub-tropical KwaZulu-Natal and consists of shallow and deep water fisheries (Fennessy and Groeneveld, 1997). The shallow water fishery, which operates at depths of 20-50 m, is situated on the Tugela Bank trawling grounds off the KwaZulu-Natal north coast and the deep water fishery operates at 100-600 m. There are limited possibilities for penaeid prawn trawling in South Africa because of the narrow width of the continental shelf along most of the South African coastline, and the Tugela Bank grounds are one of

the few suitable areas because the shelf is at it's widest here (Figure 1; Fennessy *et al.*, 1994; Fennessy, 1995). The Tugela Bank coincides with an area of mud deposition from discharges of many rivers in the area, which makes it ideal for penaeid prawns (McCormick *et al.*, 1992). The Tugela Bank supports a seasonal fishery which catches white prawn, *Penaeus indicus*, but also catches a large amount of teleosts, elasmobranchs, gastropods, cephalopods and other crustaceans as bycatch (Fennessy, 1994b; Fennessy *et al.*, 1994; Fennessy and Groeneveld, 1997).

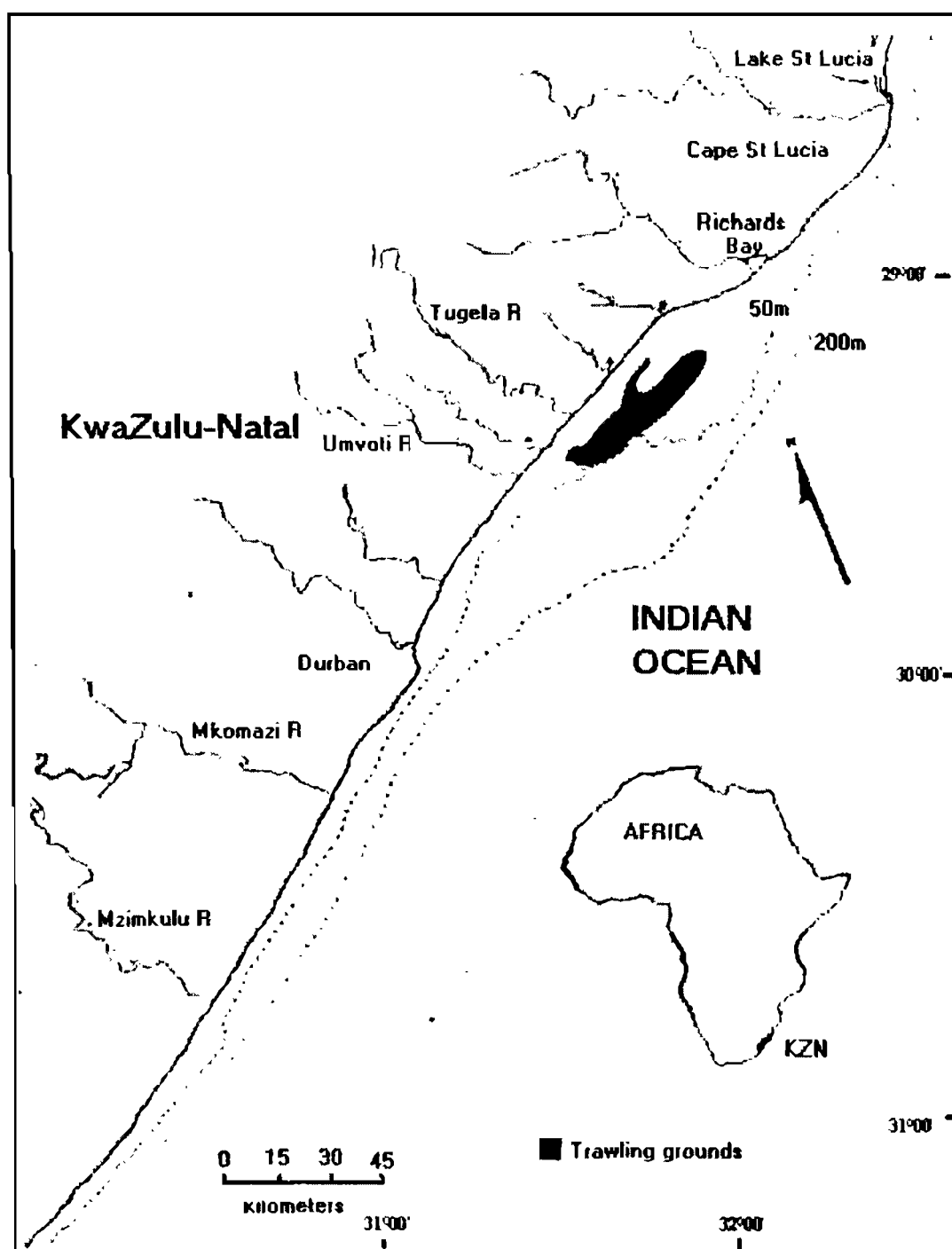
Although the scale of this fishery is small compared to other larger commercial fisheries in South Africa, the bycatch and discard proportions are high in comparison with other South African fisheries (Japp, 1997). There are other important fisheries operating in the Tugela area, such as the skiboat linefishery and the shore linefishery which are either commercial or recreational in nature (Mann *et al.* 1997). Participants in these sectors believe that the trawl fishery is responsible for the decline in species targeted by the other fisheries operating in the area (Fennessy, 1993, Mann *et al.*, 1997). The aim of this study is to investigate the species composition and to quantity of the bycatch from prawn trawlers operating on the Tugela Bank. An initial survey from 1989 to 1992 provided a baseline description of the bycatch (Fennessy *et al.*, 1994; Fennessy, 1993), and the intention of the current study was to repeat the survey after a 13-year period, to attempt to assess the impact that trawlers may have had on the region. This study also aims to provide management recommendations to reduce bycatch.

2. STUDY AREA: DESCRIPTION OF THE TUGELA BANK

The trawl grounds are situated two to four nautical miles offshore from the Matikulu River mouth (29°06'36''S, 31°36'48''E) on the KZN north coast, about 100 km from Durban (Figure 1). The grounds cover an area of approximately 500 km², with trawl depths ranging from 20 to 50 m (Fennessy, 1995). The continental shelf is at it's widest here, with the shelf break occurring 50 km offshore due to the gentle Natal Bight found between Durban and St Lucia (Lutjeharms *et al.*, 1989). The hydrodynamics of this area are dominated by wave action, rather than the Agulhas Current, which is forced further offshore by the Bight (Flemming and Hay, 1988). The south flowing Agulhas Current is a dominating dynamic feature that affects the flow along the entire KZN coastal region and it flows in a south-southeasterly direction (Schumann, 1982). There is also a strong counter-current inshore of the Agulhas Current, which flows in a north-easterly direction (Lutjeharms and Connell, 1989). This counter-current is thought to be driven by a vortex shedding from the Natal Bight, known as the Natal Pulse. These pulses originate at the Bight and grow laterally, moving downstream with average diameters of 170 km, whereas the Agulhas Current seldom exceeds 15 km in width (Lutjeharms and Connell, 1989).

The KwaZulu-Natal shelf water temperatures are generally well mixed because of the high-energy coastline (Schumann, 1982). Monthly mean sea surface temperatures measured close to the Tugela Bank ranged from 20 to 24.4°C (Fennessy, 1993). The water temperature measured at 10 m depth on the Bank was 17.75°C which was of 3°C lower than ambient shelf waters, and 8°C lower than the Agulhas Current surface waters (Lutjeharms *et al.*, 1989). The occurrence of cooler waters on the Bank is assumed to be a result of upwelling activity (Lutjeharms *et al.*, 1989), and re-circulation in the area (Schumann, 1982). In summer temperatures in the centre of the Bight are 18°C with surface temperature exceeding 22°C (Lutjeharms *et al.*, 1989). North-easterly and southerly winds predominate and they can attain speeds of up to 35 and 60 knots respectively (Hunter, 1988); these contribute to the high energies of the KwaZulu-Natal coast.

Figure 1: Location of the Tugela Bank prawn trawling grounds (After Fennessy, 1993). R=River.



3. DESCRIPTION OF THE TUGELA PRAWN FISHERY

Trawling in this fishery started around 1921, during the Fisheries and Marine Biological Survey on the S.S. Pickle, which was commissioned by the Fisheries Survey Committee of the Department of Mines and Industries (Fennessy, 1993). Trawling then occurred sporadically until 1976 when two vessels began more regular trawling (Fennessy, 1993). However, recording of catch and effort data only started in 1985 using a standard logbook system instituted by the Department of Sea Fisheries (Fennessy and Groeneveld, 1997). Regular statistics were only collected from 1988, when at the time, the fishery was comprised of 12 companies and 21 vessels (Fennessy and Groeneveld, 1997). By 1995, the number of trawlers in operation had dropped to only 7 vessels owned by 4 companies (Fennessy and Groeneveld, 1997), while only 3 trawlers were operating in 2004.

Trawling on the Tugela Bank is seasonal, occurring mainly between March and September (Fennessy, 1993; Fennessy *et al.*, 1994), and is closed from November to February, to reduce bycatch of squaretail kob, *Argyrosomus thorpei* (DEAT, 2002). The target species for this fishery is the white prawn, *F. indicus*, but there are catches of brown prawns, *Metapenaeus monoceros*, and tiger prawns, *Penaeus monodon* (Fennessy, 1993; Fennessy *et al.*, 1994). Although Total Allowable Effort (TAE) is controlled by a permit system, there have been no stock assessments done for any of the targeted species for this fishery, but it is believed to be fully exploited (DEAT, 2002).

The vessels used are steel boats with lengths ranging from 24 to 40 m, and generating between 500 and 1000 hp (Fennessy and Groeneveld, 1997). The gear used is the otter trawl, ranging from single nets to boom-operated twin or triple nets, with footrope lengths of 25 to about 70 m (Fennessy and Groeneveld, 1997). The stretched mesh size is 55 mm in both the wings and the codend (Fennessy and Groeneveld, 1997). Trawling occurs on a 24 h basis with trawl speeds of two to three knots and drag duration averaging 4-5 hours. Trawlers carry crews of between 12 and 20, (Fennessy and Groeneveld, 1997) and the duration of trawl trips can be up to one month.

At the end of each trawl the catch is sorted by the crew and the prawn component is graded by size, packed in cartons and blast frozen as soon as possible (Fennessy, 1993). A large portion of the retained catch is prawns but the largest portion of the catch is bycatch, which, if commercially viable, is packed for marketing. This includes crabs, *Portunus sanguinolentus*, cephalopods, *Uroteuthis duvauceli*, *Sepiella cyanea* and several marketable fish species (Fennessy, 1993). The rest of the bycatch, which has a low market value is discarded.

4. MATERIALS AND METHODS

4.1 Sampling procedure

An initial survey of the bycatch species composition of the prawn trawlers operating on the Tugela Bank was done between May 1989-June 1992 (Fennessy, 1993) and the current study between March 2003 and July 2004 was a follow up to this initial survey.

An observer from Marine and Coastal Management (MCM), a directorate of the South African Department of Environmental Affairs and Tourism, was placed onboard the trawlers to collect discarded bycatch samples. There were a maximum of 120 observer days per year allocated for both shallow and deep water trawling. The trawler trips were of 3-4 weeks duration and trawlers sometimes switched from shallow to deep water trawling during a trip depending on the amount of target species available, so there was no control over the regularity of sampling. Four drags were sampled each day, three of which were processed onboard the trawler by the observer and one sample was frozen for on-shore processing. After the crew had removed the target species and the retained bycatch, the observer removed larger elasmobranchs from the bycatch and collected a +20 kg sample of the discarded catch. A species identification catalogue was provided to assist with onboard species identification, and the species that were not included in the catalogue were labelled with a name and trawl details for later identification. Both the observer and the skipper recorded depth, position, time of day and duration of each sampled trawl in drag books. The quantities of retained prawn, retained bycatch and discarded bycatch were also recorded by the observer.

The frozen samples of discarded bycatch were taken to the Oceanographic Research Institute (ORI), in Durban for processing and analysis on-shore. Sorting of the samples included a breakdown of the sample by weight and number of all taxa, and recording the total lengths (TL) of at least fifty individuals of the three most commonly recorded teleost species. A third of the crustacean component was sub-sampled when it was too large to count and weigh each individual. Afterwards necessary corrections were made to estimate actual weight and number of crustaceans in that particular sample. Smith and Heemstra (1986) was used for teleosts species identification and Barnard (1950) for most of the crustaceans. Unidentified specimens were sent to the South African Institute of Aquatic Biodiversity (SAIAB) in Grahamstown and IZIKO Natural History Museum in Cape Town for positive identification. The data were captured onto Microsoft Excel spreadsheets, from which tables of the catch composition were produced illustrating the occurrence and relative contributions by weight and number of the teleosts, invertebrates and elasmobranchs to the total bycatch.

4.2 Analysis of data

Much of the data is descriptive, and is presented in the form of comparative tables. A more rigorous approach, such as multivariate analysis, was considered and rejected, since there was little control over sampling, which was ad hoc (see results).

4.2.1 Total discarded bycatch and retained bycatch

The total annual discarded bycatch for 2003 was estimated by scaling up discarded catch rates (kg/hr) using total fleet effort. The fleet effort, weights of the target species and retained bycatch were obtained from information recorded in drag books by skippers, which were submitted to Marine and Coastal Management. Only 2003 information was available at the time of the analysis.

4.2.2 Differences in discard rates

The effects of gear size, season, depth and day or night on discard rates (kg discard.trawl hour⁻¹) were tested using t-tests. These differences were tested for all the trawls for which the observer had recorded discard rates in the 2003-2004 study period.

4.2.3 Length frequencies

The total lengths of the three most commonly recorded and consistently discarded bycatch species (*O. ruber*, *J. fuscolineatus* and *J. dorsalis*) were measured using a measuring board. A maximum of 50 randomly selected individuals of each species were measured per sample. The data were captured onto Excel worksheets and length frequency distribution histograms were produced. The Kolmogorov-Smirnov two sample test was applied to the data sets from 1989-1992 and 2003-2004 to test for any differences.

5. RESULTS

5.1 General

A total of 47 on-shore and 121 onboard discarded bycatch samples were collected between April 2003 and July 2004. The trawling trips were approximately one month in duration. The observer was instructed to collect and sort three onboard samples per day, but this did not always happen because of bad weather. There were also difficulties experienced in obtaining monthly samples due to low catch rates of the white prawn, *F. indicus*, which resulted in reduced fishing effort (Table 1.1).

Table 1.1: The number of discarded on-shore and onboard bycatch samples collected from Tugela Bank prawn trawlers during 2003-2004.

Month	Year	On-shore samples	Onboard samples
April	2003	7	20
May	2003	13	24
March	2004	10	21
April	2004	2	6
June	2004	2	3
July	2004	13	47
TOTAL		47	121

For analysis purposes, two seasons were defined based on sea surface temperatures: a 'warm' season (December-April) and a 'cool' season (May-November). Trawling occurred at depths ranging between 16 and 50 m and trawls were categorised as 'shallow' (16-24 m), 'medium' (26-30 m) and 'deep' (33-45 m). These depth and season categories (Table 1.2) were selected based on those described in Fennessy *et al.* (1994), to facilitate comparison with the 1989-1992 data. For convenience, the 1989-1992 data will be referred to as the 1990s data set and the 2003-2004 as the 2000s data set.

Table 1.2: The number of discarded on-shore and onboard bycatch samples collected from Tugela Bank prawn trawlers at different depths, seasons and time of day (April 2003-July 2004). On-shore = 47, onboard = 121.

Season	On-shore	Onboard	Depth	On-shore	Onboard	Time of day	On-shore	Onboard
Cool	28	74	Shallow	40	100	Day	41	105
Warm	19	47	Medium	7	16	Night	6	16
			Deep		5			

5.2 Overall catch composition of discarded bycatch

A total discarded on-shore bycatch sample weight of 842 kg, comprising 23 401 vertebrate and invertebrate individuals, was obtained during the survey. There were 14 560 teleost individuals and 8 812 invertebrates, making up 62% and 38% by number respectively of all samples. The onboard samples comprised a total of 43 567 individuals, 29 163 (67%) of which were teleosts and 12 276 (28%) were invertebrates.

Only 24 elasmobranch (sharks and rays) individuals were found in the on-shore samples whereas the onboard samples comprised 2 104 elasmobranchs (Table 2.1). This is because the observer was instructed to exclude the larger elasmobranchs from collected samples. The top five elasmobranch species in the onboard samples were (scalloped hammerhead shark) *Sphyrna lewini*, (duckbill ray) *Pteromyiaeus bovinus*, (spotted eagle ray) *Aetobatus narinari*, (stingray) *Himantura gerrardi* and (butterfly ray) *Gymnura natalensis* (Table 2.1). The identity of the onboard elasmobranchs could not be verified and should be treated with caution.

Table 2.1: Overall relative abundance of discarded bycatch elasmobranch species from Tugela Bank commercial prawn trawls (April 2003-July 2004). On-shore samples = 47, onboard samples = 121.

Scientific name	On-shore	Onboard	Onboard	Onboard
Species	Total number	Total number	Percent by number	Percent of trawls
<i>Sphyrna lewini</i>	2	534	25.4	63.2
<i>Pteromyiaeus bovinus</i>	0	463	22.0	51.6
<i>Aetobatus narinari</i>	0	361	17.1	18.3
<i>Himantura gerrardi</i>	5	197	9.4	44.7
<i>Gymnura natalensis</i>	0	134	6.4	55.9
<i>Rhinobatos annulatus</i>	10	126	6.0	37.1
<i>Rhinoptera javanica</i>	0	94	4.5	4.6
<i>Dasyatis chrysomata</i>	1	39	1.9	26.7
<i>Rhinobatos leucospilus</i>	0	28	1.3	5.6
<i>Squatina africana</i>	1	28	1.3	13.4
<i>Scylliogaleus queketti</i>	0	25	1.2	5.1
<i>Torpedo sinuspersici</i>	4	21	1.0	3.2
<i>Himantura varna</i>	0	11	0.5	19.7
<i>Rhynchobatus djiddensis</i>	0	10	0.5	10.6
<i>Mustelus mosís</i>	0	9	0.4	1.3
<i>Dasyatis thetidis</i>	0	7	0.3	21.2
<i>Rhinobatos holerhynchus</i>	0	5	0.2	3.1
<i>Himantura thrao</i>	0	4	0.2	1.1
<i>Carcharhinus obscurus</i>	0	3	0.1	2.2
<i>Rhizoprionodon acutus</i>	0	3	0.1	2.2
<i>Taeniura melanospilos</i>	0	2	0.1	1.2
<i>Holohalaelurus punctatus</i>	1	0	0	0

The top five most commonly recorded invertebrate species made up 74% and 87% of all invertebrate on-shore and onboard samples by number respectively, with portunid crabs dominating the invertebrate catch. In the on-shore samples these were (crabs) *Portunus hastatoides*, *P. sanguinolentus*, *Matuta* sp.,

(mantis shrimp) *Squilla armata* and (cuttlefish) *Sepiella cyanea*. The top five invertebrate species in the onboard samples were *Matuta* sp., *P. sanguinolentus*, *S. armata*, *P. hastatoides* and (cuttlefish) *Sepia vermiculata* (Table 2.2). There is, however a likelihood of misidentification of the onboard invertebrate species because, unlike for the teleosts, the observer did not have an extensive identification catalogue for this group. For example, the considerable numbers of *S. vermiculata* recorded in the onboard samples compared to on-shore samples could be because of misidentification by the observer, as this species can be easily confused with *S. cyanea* (Table 2.2). There was also a high number of unidentified decapod crustaceans (crabs) recorded in the onboard samples because of difficulty in species identification.

Table 2.2: Overall relative abundance of discarded bycatch invertebrate species from Tugela Bank commercial prawn trawls (April 2003-July 2004). On-shore samples = 47, onboard samples = 121.

Scientific Name	On-shore	Onboard	On-shore	Onboard	Onshore/ Onboard	On-shore	Onboard
Species	Total number	Total number	Percent by number	Percent by number	Percent difference	Percent of trawls	Percent of trawls
<i>Portunus hastatoides</i>	1720	928	19.5	7.5	12	85.1	16.5
<i>Portunus sanguinolentus</i>	1437	3596	16.3	29.2	12.9	93.6	93.4
<i>Matuta</i> sp.	1420	3748	16.1	30.5	14.4	76.6	80.2
<i>Squilla armata</i>	1367	1719	15.5	14	1.5	78.7	62.8
<i>Sepiella cyanea</i>	537	0	6.1	0	6.1	95.7	0
<i>Fenneropenaeus indicus</i>	507	0	5.8	0	5.8	68.1	0
<i>Mactra aequisulcata</i>	456	0	5.2	0	5.2	63.8	0
<i>Parapenaeopsis acclivirostris</i>	376	230	4.3	1.9	2.4	66	14.9
<i>Charybdis merguensis</i>	188	3	2.1	0	2.1	40.4	2.5
<i>Anadara natalensis</i>	137	0	1.6	0	1.6	42.6	0
<i>Ficus ficus</i>	112	0	1.3	0	1.3	46.8	0
<i>Chlamys furtoni</i>	101	0	1.2	0	1.2	44.7	0
<i>Conchoecetes artificiosus</i>	100	187	1.1	1.5	0.4	48.9	30.6
<i>Uroteuthis duvauceli</i>	76	0	0.9	0	0.9	53.2	0
<i>Philyra globulosa</i>	48	178	0.5	1.4	0.9	42.6	17.4
<i>Bufonaria crumena</i>	35	0	0.4	0	0.4	36.2	0
<i>Gonionepuntus africanus</i>	31	7	0.4	0.1	0.3	25.5	2.5
<i>Platylambrus quemvis</i>	23	12	0.3	0.1	0.2	25.5	5.8
<i>Polynices didyma</i>	22	0	0.2	0	0.2	19.1	0
<i>Trachypenaeus curvirostris</i>	20	0	0.2	0	0.2	8.5	0
<i>Eudolium pyriforme</i>	13	0	0.1	0	0.1	19.1	0
<i>Ixoides cornutus</i>	12	0	0.1	0	0.1	17	0
<i>Charybdis cruciata</i>	12	0	0.1	0	0.1	21.3	0
Unidentified bivalves	12	0	0.1	0	0.1	4.3	0
Unidentified sea hare	12	0	0.1	0	0.1	12.8	0
<i>Panulirus homarus</i>	11	37	0.1	0.3	0.2	16.5	16.5
Unidentified asteroids	10	19	0.1	0.2	0	19.1	5.8
<i>Dorippe lanata</i>	7	0	0.1	0	0.1	12.8	0
<i>Doclea muricata</i>	5	0	0.1	0	0.1	4.3	0
<i>Charybdis natator</i>	4	0	0.1	0	0.1	8.5	0
<i>Sepia vermiculata</i>	1	757	0	6.2	6.1	2.1	38.0
Unidentified jellyfish	0	211	0	1.7	1.7	0	17.4
<i>Loligo duvauceli</i>	0	60	0	0.5	0.5	0	18.2
Unidentified cephalopod	0	41	0	0.3	0.3	0	5.8
Unidentified decapod crustaceans	0	543	0	4.4	4.4	0	11.6

The top five teleost species made up 64% and 66% by number of the total teleost catch of the on-shore and onboard discarded bycatch samples respectively. The most commonly recorded fish species,

Otolithes ruber, *Johnius dorsalis*, *J. fuscolineatus*, *Secutor ruconius* and *Trichiurus lepturus*, occurred in similar proportions in both sample sets (Table 2.3).

Table 2.3: Overall relative abundance of discarded bycatch teleost species from Tugela Bank commercial prawn trawls (April 2003-July 2004). Only species that contributed greater than or equal to 0.1% by number in the on-shore or onboard samples were included. A comprehensive list is provided in Appendix 1. On-shore n= 47, onboard n=121.

Scientific Name	On-shore	Onboard	On-shore	Onboard	On-shore/ Onboard	On-shore	Onboard
Species	Total number	Total number	Percent by number	Percent by number	Percent difference	Percent of trawls	Percent of trawls
<i>Otolithes ruber</i>	2904	6926	20	23.7	3.8	95.7	94.3
<i>Johnius dorsalis</i>	2337	5617	16.1	19.3	3.2	100	87.8
<i>Johnius fuscolineatus</i>	1875	1960	12.9	6.7	6.2	100	57.7
<i>Secutor ruconius</i>	1303	2482	9.0	8.5	0.4	63.8	48
<i>Trichiurus lepturus</i>	887	2396	6.1	8.2	2.1	83.0	82.1
<i>Arobus nibe</i>	512	688	3.5	2.4	1.2	61.7	41
<i>Thryssa vitirostris</i>	479	955	3.3	3.3	0	83.0	75.6
<i>Pomatomus saltatrix</i>	412	916	2.8	3.1	0.3	78.7	69.9
<i>Pellona ditchela</i>	411	740	2.8	2.5	0.3	55.3	18.7
<i>Leiognathus equula</i>	360	248	2.5	0.9	1.6	40.4	13.8
<i>Upeneus vittatus</i>	342	654	2.4	2.2	0.1	48.9	41.5
<i>Alepes djedaba</i>	295	357	2.0	1.2	0.8	44.7	17.1
<i>Polydactylus malagasyensis</i>	292	299	2.0	1.0	1.0	68.1	30.1
<i>Cynoglossus attenuatus</i>	227	900	1.6	3.1	1.5	85.1	75.6
<i>Carangoides malabaricus</i>	180	289	1.2	1.0	0.2	42.6	8.9
<i>Cociella heemstrai</i>	180	363	1.2	1.2	0	70.2	71.5
<i>Saurida undosquamis</i>	178	448	1.2	1.5	0.3	57.0	59.3
<i>Parachannaichthys polynema</i>	151	0	1.0	0	1.0	27.7	0
<i>Pomadoury olivaceum</i>	139	371	1.0	1.3	0.3	38.3	28.5
<i>Argyrosomus thorpei</i>	117	80	0.8	0.3	0.5	40.4	28.1
<i>Stolephorus punctifer</i>	113	29	0.8	0.1	0.7	21.3	9.8
<i>Sardinops sagax</i>	112	925	0.8	3.2	2.4	10.6	18.7
<i>Cynoglossus lida</i>	98	48	0.7	0.2	0.5	40.4	10.6
<i>Gazza minuta</i>	91	328	0.6	1.1	0.5	38.3	25.2
<i>Lagocephalus guentheri</i>	59	76	0.4	0.3	0.1	46.8	19.5
<i>Sphyrna acutipinnis</i>	58	54	0.4	0.2	0.2	21.3	13.0
<i>Drepane longimanus</i>	50	126	0.3	0.4	0.1	42.6	38.2
<i>Parastromateus niger</i>	33	68	0.2	0.3	0	19.1	14.0
<i>Paralichthodes algoensis</i>	28	79	0.2	0.3	0.1	38.3	20.3
<i>Cyclichthys orbicularis</i>	26	33	0.2	0.1	0.1	27.7	9.8
<i>Lagocephalus inermis</i>	14	47	0.1	0.2	0.1	21.3	17.1
<i>Ariomma indica</i>	13	37	0.1	0.1	0	17	4.1
<i>Epinephelus andersoni</i>	12	43	0.1	0.1	0.1	17	14.6
<i>Solen bleekeri</i>	9	64	0.1	0.2	0.2	4.3	9.8
<i>Champsodon capensis</i>	0	171	0	0.6	0.6	0	23.6
Unidentified fish 1	0	95	0	0.3	0.3	0	2.4
<i>Pomadoury kaakan</i>	2	49	<0.1	0.4	0.4	2.1	8.0
<i>Pomadoury commersoni</i>	0	46	0	0.2	0.2	0	3.0
Unidentified fish 2	0	35	0	0.1	0.1	0	3.3
<i>Uroconger lepturus</i>	0	17	0	0.1	0.1	0	8.9

However, the relatively higher contribution of *J. dorsalis* in the onboard samples as compared to the on-shore samples is possibly because of confusion with *J. fuscolineatus* and probably also explains the

discrepancy between on-shore and onboard numbers of *J. fuscilineatus*. In general, though there were only minor discrepancies between onboard and on-shore samples in terms of relative abundance. In terms of percent abundance, however, there were some substantial discrepancies between the two data sets. For example, although the relative contributions of *A. nibe* and *Carangoides malabaricus* were similar in both sets of data, the percent of trawls occurrence in the onboard samples was much lower than in the on-shore samples (Table 2.3). These discrepancies could be due to onboard identification problems, or due to onboard sampling coinciding with the trawling of large, irregularly occurring shoals of these species.

5.2.1 Comparison of data from 1989-1992 and 2003-2004

Comparisons were made between the 1989-1992 (1990s) data and the 2003-2004 (2000s) on-shore data, to identify any changes in the discarded bycatch species composition of the Tugela prawn trawlers. Only the fish data were used for comparison because the 1989-1992 invertebrate data were not available. To reduce confounding effects as much as possible, only samples collected during the same months and depth ranges were included in the analysis. A total of 58 samples from 1989-1992 were selected for comparison (Table 2.4). Season and depth categories were based on those in Fennessy *et al.* (1994).

Table 2.4: The number of discarded on-shore samples collected from Tugela Bank prawn trawlers at different depths, seasons and time of day (1989-1992, n=58).

Season	Number of samples	Depth	Number of samples	Time of day	
Cool	42	Shallow	22	Day	45
Warm	16	Medium	31	Night	13
		Deep	5		

Overall species composition in both surveys was similar but contributions by individual species were different (Table 2.5). The top five most commonly recorded species in the 1990s survey were *J. dorsalis*, *J. fuscilineatus*, *O. ruber*, *T. lepturus* and *A. nibe*, whereas in 2000s, *O. ruber*, *J. dorsalis*, *J. fuscilineatus*, *S. ruconius* and *T. lepturus* were the most common (Table 2.5).

Of the 15 most commonly recorded fish species in the 1990s, 11 showed a decline in relative abundance of 5-91%, nine of which are demersals, and four species showed an increase of 20-189%. Of the 11 species which declined, nine are demersal, one is pelagic and one (*Trichiurus lepturus*) is benthopelagic. Of the four species which increased in relative abundance, two are demersal and two are pelagic. The most marked increase was shown by three pelagic species *S. ruconius*, *P. saltatrix* and *L. equula*, which increased in relative abundance by more than 2000%. As a group, members of the family Sciaenidae (kobs) showed an overall decline in their contribution to relative abundance by number from 63.6% in the 1990s to 53.3% in the 2000s. Three sciaenid species (*J. dorsalis*, *J. fuscilineatus* and *A.*

libe) showed a decline in percent relative abundance of 26-59%, while *O. ruber* and *A. thorpei* increased in relative abundance by 32% and 189% respectively.

Examinations of changes in percent frequency of occurrence show that there are considerable differences between the two data sets for some species. Ten of the 15 most commonly recorded species in the 1990s showed a decline of 3-68% in percent occurrence in the 2000s, and eight of these are demersal species. Only one pelagic species, *T. vitreosus*, showed a decline in frequency of occurrence. Five species (*Parastromateus niger*, *Epinephelus andersoni*, *L. equula*, *P. saltatrix* and *Pseudorhombus natalensis*) showed an increase of more than 200% in frequency of occurrence and three of these are pelagic species.

There were 24 species in the 1990s samples which did not occur in the 2000s, some of which were quite commonly recorded in the earlier samples, e.g. *B. maclellandi* and *U. lepturus* (Table 2.5). Similarly, there were 25 species in the 2000s samples which did not occur in the 1990s, some of which were also commonly recorded in the current samples, e.g. *A. djedaba* and *Stolephorus punctifer* (Table 2.5). However, 10 of these species were recorded in other samples from the 1990s, i.e. from months and/depths that were not included in the comparison (see Appendix).

Table 2.5:

Comparison of overall relative abundance of teleost species between 1989-1992 (n=58) and 2003-2004 (on-shore samples = 47) surveys. Only species that contributed greater than or equal to 0.1% by number were included. + or - signs indicate an increase or decrease in percent differences in relative abundance and frequency of occurrence. The percent differences in relative abundance and frequency of occurrence values have been rounded off for better presentation.

Scientific name	1989-1992			2003-2004			Percent Change in abundance	Percent Change in frequency of occurrence	Pelagic (P) Demersal (D)
	Number	Percent number	Percent of trawls	Number	Percent number	Percent of trawls			
<i>Johnius dorsalis</i>	5237	22.0	96.6	2337	16.1	100	-27	+4	D
<i>Johnius fuscolineatus</i>	4167	17.5	98.3	1875	12.9	100	-26	+2	D
<i>Oriolithes ruber</i>	3617	15.2	98.3	2904	20.0	95.7	+32	-3	D
<i>Trichiurus lepturus</i>	2456	10.3	98.3	887	6.1	83.0	-41	-16	P/D
<i>Atrobuca nibe</i>	2041	8.6	93.1	512	3.5	61.7	-59	-34	D
<i>Thryssa vitirostris</i>	1752	7.3	98.3	479	3.3	83.0	-55	-16	P
<i>Lagocephalus guentheri</i>	840	3.5	84.5	59	0.4	46.8	-89	-45	D
<i>Drepane longimanus</i>	797	3.3	82.8	50	0.3	42.6	-91	-49	D
<i>Cynoglossus attenuatus</i>	495	2.1	94.8	227	1.6	85.1	-23	-10	D
<i>Cynoglossus lida</i>	271	1.1	58.6	98	0.7	40.4	-38	-31	D
<i>Carangoides malabaricus</i>	239	1.0	39.7	180	1.2	42.6	+20	+7	P
<i>Pellona ditchela</i>	231	1.0	37.9	411	2.8	55.3	+189	+46	P
<i>Gazza minuta</i>	150	0.6	46.6	91	0.6	38.3	-5	-18	D
<i>Pomadasyd olivaceum</i>	145	0.6	36.2	139	1.0	38.3	+65	+6	D
<i>Lagocephalus inermis</i>	135	0.6	67.2	14	0.1	21.3	-82	-68	D
<i>Saurida undosquamis</i>	117	0.5	51.7	178	1.2	57.4	+145	+11	D
<i>Upeneus vittatus</i>	115	0.5	37.9	342	2.4	48.9	+398	+29	D
<i>Polydactylus malagasyensis</i>	111	0.5	69.0	292	2.0	68.1	+330	-1	D
<i>Arionomma indica</i>	103	0.4	36.2	13	0.1	17.0	-77	-53	P
<i>Secutor ruconius</i>	95	0.4	36.2	1303	9.0	63.8	+2160	+76	P
<i>Cociella heemstrai</i>	72	0.3	46.6	180	1.2	70.2	+298	+51	D
<i>Solen bleekeri</i>	71	0.3	50.0	9	0.1	4.3	-66	-91	D
<i>Argyrosomus thorpei</i>	66	0.3	20.7	117	0.8	40.4	+189	+95	D
<i>Thryssa setirostris</i>	66	0.3	31.0	32	0.2	29.8	-20	-4	P
<i>Bregmaceros maclellandi</i>	57	0.2	15.5	0	0	0	-100	-100	P
<i>Uroconger lepturus</i>	51	0.2	17.2	0	0	0	-100	-100	D
<i>Sardinops sagax</i>	38	0.2	10.3	112	0.8	10.6	+285	+3	P
<i>Sardinella albella</i>	31	0.1	12.1	0	0	0	-100	-100	P
<i>Antennarius striatus</i>	26	0.1	25.9	0	0	0	-100	-100	D
<i>Epinephelus andersoni</i>	25	0.1	3.4	12	0.1	17.0	-5	+400	D
<i>Mene maculata</i>	24	0.1	5.2	0	0	0	-100	-100	P
<i>Parachaeturichthys polynema</i>	17	<0.1	15.5	151	1.0	27.7	+940	+79	D
<i>Aesopia cornuta</i>	16	<0.1	15.5	0	0	0	-100	-100	D
<i>Pseudorhombus natalensis</i>	13	<0.1	6.9	43	0.3	21.3	+200	+209	D
<i>Cyclichthys orbicularis</i>	11	<0.1	10.3	26	0.2	27.7	+100	+169	D
<i>Dysomma anguillare</i>	9	<0.1	12.1	11	0.1	19.1	-20	+58	D
<i>Leiognathus equula</i>	8	<0.1	8.6	360	2.5	40.4	+2370	+370	P
<i>Pomadasyd maculatum</i>	8	<0.1	6.9	13	0.1	8.5	-10	+23	D
<i>Callionymus marleyi</i>	6	<0.1	6.9	20	0.1	8.5	+40	+23	D
<i>Parastromateus niger</i>	2	<0.1	3.4	33	0.2	19.1	+100	+462	P
<i>Alepes djedaba</i>	0	0	0	295	2.0	44.7	~	~	P
<i>Archamia mozambiquensis</i>	0	0	0	39	0.3	23.4	~	~	D
<i>Pagellus bellottii natalensis</i>	0	0	0	31	0.2	2.1	~	~	D
<i>Paralichthodes algoensis</i>	0	0	0	28	0.2	38.3	~	~	D
<i>Pseudorhombus natalensis</i>	0	0	0	43	0.3	21.3	~	~	D
<i>Sillago sihama</i>	0	0	0	11	0.1	17.0	~	~	D
<i>Sphyræna acutipinnis</i>	0	0	0	58	0.4	21.3	~	~	P
<i>Sphyræna flavicauda</i>	0	0	0	12	0.1	12.8	~	~	P
<i>Stolephorus punctifer</i>	0	0	0	113	0.8	21.3	~	~	P

5.3 Differences in discard rates

The effects of gear size (Table 3.1), season, depth and day or night on discard rates (kg discard.trawl hour⁻¹) were tested using t-tests. These differences were tested for all the trawls for which the observer had recorded discard rates in the 2000s study period.

Table 3.1: Gear types used in prawn trawlers on the Tugela Bank during the 2000s study and their respective size coding.

Gear type	Size coding	Footrope length(m)
Florida Flyer (FF)	Medium	2x24
Tri-net (2)	Medium	2x24
4-seam	Medium	47
Chico 500	Large	70
Tri-net (3)	Large	3x24

When testing for the effect of a factor on discard rate, the potentially confounding effect of other factors was minimised. For example, to test for the discard rate differences between day and night trawls, only trawls done during the cool season, in shallow depths and using large gear were considered. The combination of factors was selected to maximise the number of trawls being tested (Table 3.2).

Table 3.2: Number of Tugela Bank trawls for which discard rates were estimated, with associated variables.

Day/ Night	Number of trawls	Season	Number of trawls	Depth	Number of trawls	Gear size	Number of trawls
Day	105	Cool	74	Shallow	100	Medium	55
Night	16	Warm	47	Medium	16	Large	66
				Deep	5		
Total	121						

There was no difference in discard rate between trawls from shallow and medium depths. The test did not include trawls from deep depths because there were only five trawls in deep waters. However, the t-tests done on the gear size and season factors revealed significant differences in

5.4 Length frequencies

The mean lengths for *O. ruber*, *J. dorsalis* and *J. fuscolineatus* in this study were higher than those found in the 1989-1992 study by Fennessy (1993), suggesting that the fish caught are larger now than 10 years ago (Table 5.1). The Kolmogorov-Smirnov tests revealed that there were significant differences between the 1990s and 2000s length frequency distributions of all three species (Table 5.2). However, the interpretation of these results is complicated by the fact that the mesh size has increased substantially since the 1990s study, from 38 mm to 55 mm (stretched mesh).

Table 4.1: A comparison of the total lengths of three trawled sciaenid species in trawler discards from 1989-1992 (1990s) and 2003-2004 (2000s).

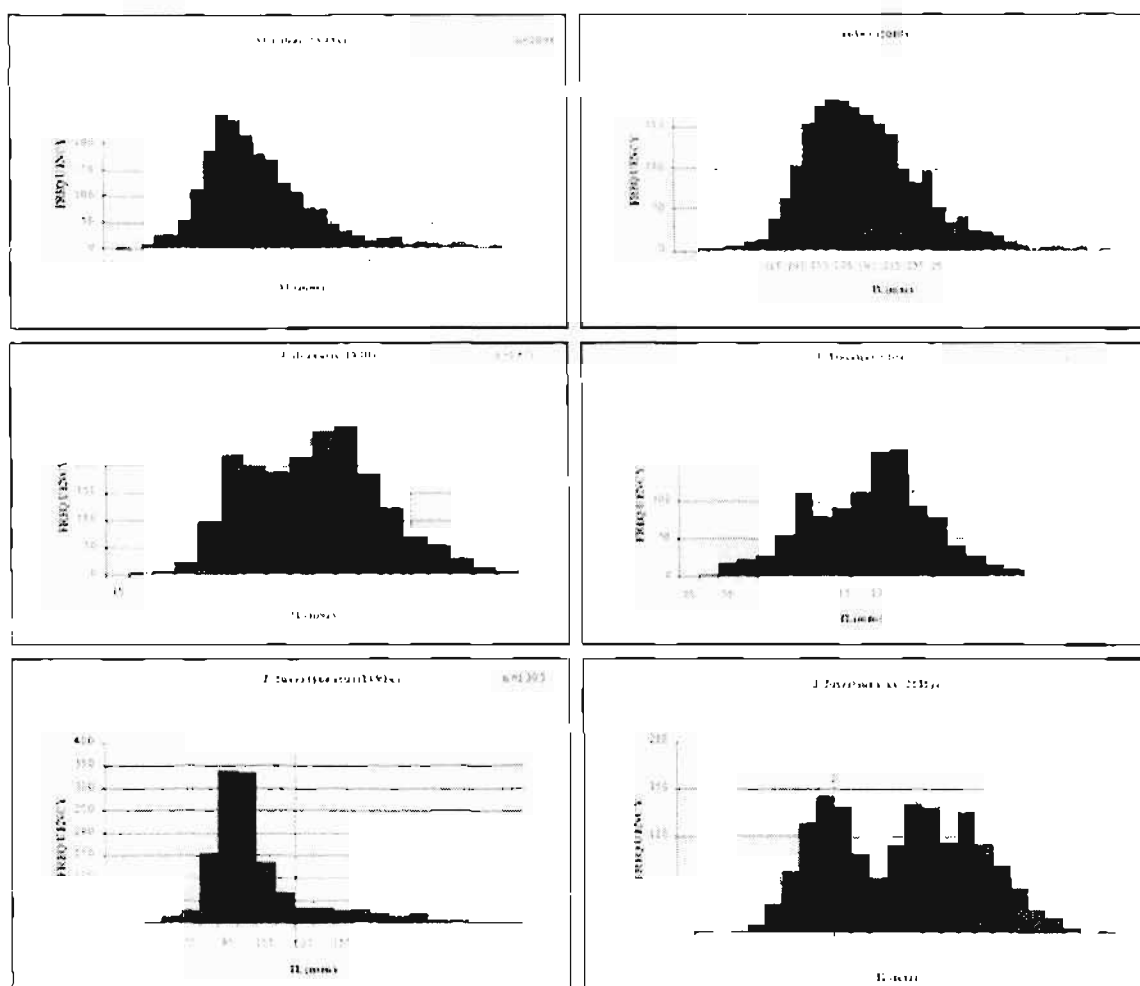
	<i>O. ruber</i>		<i>J. fuscolineatus</i>		<i>J. dorsalis</i>	
	1990s	2000s	1990s	2000s	1990s	2000s
N	2096	2067	1303	1459	1978	1123
MEAN	151	187.8	107	154.3	116	126
RANGE (TL mm)	39-456	50-390	52-248	45-275	38-220	45-205

Table 4.2: Results of the Kolmogorov-Smirnov test to illustrate differences between the 1990s and 2000s length frequency distributions of three sciaenid species from 1989-1992 (1990s) and 2003-2004 (2000s).

	<i>O. ruber</i>	<i>J. dorsalis</i>	<i>J. fuscolineatus</i>
Critical value	0.042	0.051	0.052
Test Value (D)	0.348	0.158	0.531

The length frequency distribution for *O. ruber* was positively skewed in the 1990s, while in the 2000s, it is approaching a normal distribution because there is a higher proportion of larger fish (Figure 3). The length frequency distribution for *J. fuscolineatus* in the 1990s was unimodal, while in the 2000s, it has two modes (Figure 3). This suggests that most of the size classes of these two species are now vulnerable to trawling, whereas before, mainly smaller size classes were trawled. However, the length frequencies of *J. dorsalis* from both year data sets are normally distributed and similar in appearance (Figure 3).

Figure 2: A comparison of length frequency distributions of three sciaenid species in trawler discards from 1989-1992 (1990s) and 2003-2004 (2000s) showing size at 50% maturity.



6. DISCUSSION

6.1 General

Bycatch is defined as “the part of the gross catch which is captured incidentally to the species toward which there is directed effort” (Saila, 1983). This bycatch can either be retained or discarded depending on its marketability/use. The discarded catch is the part of the bycatch that is thrown back into the sea. In recent years, most of the world’s marine fisheries have been faced with the problem of bycatch (Broadhurst, 2000), as it is believed to play a significant role in the decline of many of the world’s fish stocks. This is because discarded bycatch makes up approximately 10.6% of the total annual global landings (Japp, 1997). Many of the species caught as bycatch are targeted by other fisheries and this additional catch reduces the recruitment, biomass and yield of stocks that form these fisheries (Vaughan *et al.*, 1991; Kennelly *et al.*, 1993; Lowry *et al.*, 2001). As a large proportion of the bycatch consists of juveniles and subadults (Andrew and Pepperell, 1992; Broadhurst, 2000), this mortality represents a threat to biodiversity and long-term sustainability of marine fisheries.

Commercial prawn trawling occurs in the majority of the world’s oceans and accounts for a significant portion of the total production of the world’s marine fisheries (Broadhurst, 2000). Most of the world’s prawn trawl fisheries are situated in shelf waters of tropical or sub-tropical waters, where there is high abundance and diversity of other organisms (van der Elst, 1988, Fennessy, 1993). This factor, coupled with the highly indiscriminate nature of the otter trawl used for prawn trawling, accounts for the high percentage of bycatch in prawn trawl fisheries (Fennessy, 1994a).

The only crustacean trawl fishery in South Africa operates on the east coast of KwaZulu-Natal (KZN), and although it is small in scale, it is of considerable local commercial importance. The shallow water prawn trawl fishery is located on the Tugela Bank, and it is associated with a high diversity of bycatch. A previous study found this fishery to have the highest bycatch and discard rates of all the South African fisheries (Japp, 1997). There is also evidence that the fishery has been partly responsible for the decline of at least one species targeted by other fisheries (Fennessy, 1994b). An initial survey from 1989-1992 of the Tugela fishery bycatch composition focused on identifying species that were caught by trawlers and other fisheries sectors. A comparison of the bycatch data from 1989-1992 with the current information permitted a preliminary assessment of the wider potential impacts of trawling on fish in the area.

6.2 Overall catch composition and effort

Prawn trawling for the white prawn, *F. indicus* occurs on the Tugela Bank, but there are also significant catches of teleosts, elasmobranchs and invertebrate species, as shown in the earlier study by Fennessy (1993). The analysis of bycatch species composition in the current study supports this, showing high

ichthyofaunal catch (122 species). Other studies have also recorded high and diverse levels of bycatch in commercial prawn trawl fisheries e.g. Australia (Harris and Poiner, 1991); Indonesia (Evans and Wahyu, 1996); America (Nance and Scott-Denton, 1997) and Scotland (Stratoudaski *et al.*, 2001).

There were 22 elasmobranch species in the discarded bycatch, which contributed about 5% to the total discarded bycatch by number. Numerically four of the top five elasmobranchs are rays, which are demersal species and therefore highly susceptible to capture in prawn trawls (Stobutzki, 2002). Three of the top five species in the onboard samples occurred in more than 50% of the trawls. The impact of the prawn trawl fishery on the sustainability of elasmobranch species has yet to be addressed. An Australian study found that more than 50% of elasmobranchs caught in prawn trawls were immature and 66% died in the trawl net (Stobutzki *et al.*, 2002). Fennessy (1994a) also found high mortalities of some elasmobranchs in prawn trawls in South Africa.

In this study the invertebrates contributed 38% and 28% to the on-shore and onboard samples respectively, with portunid crabs dominating the catch. These are benthic and therefore vulnerable to capture in prawn trawls. Three of the top five most abundant invertebrate species in the on-shore samples occurred in more than 85% of trawls, and two species in the onboard samples occurred in more than 80% of trawls. Two of the five most commonly caught invertebrate species, the spotted crab (*P. sanguinolentus*) and swimming crab (*Matuta sp.*) were also abundant in a study of prawn trawl bycatch done by Walter (1997) in the East Frisian islands.

The Tugela prawn trawl bycatch consisted mostly of teleosts, contributing 64% and 66% to the total discarded bycatch by number in both the on-shore and onboard samples respectively. The fish bycatch community comprised of a mix of demersal and pelagic species with the demersal species dominating the catch. Most of the fish are shoaling species which also accounts for their vulnerability to the trawl gear (Fennessy, 1994a). Four of the top five fish species are demersal and *T. lepturus* is a semi-demersal species. The demersal nature of prawn trawling and slow trawling speeds probably accounted for the relatively lower abundance of pelagic species. A few species dominated by number and weight, a trend reported from many other prawn trawl bycatch studies, including in the Gulf of Carpentaria, Australia (Blaber *et al.*, 1990) ; the Great Barrier Reef, Australia (Watson *et al.*, 1990) and the Gulf of Paria, Trinidad (Maharaj and Recksiek, 1991). It has also been observed in other bycatch studies that the largest component of bycatch is fish (Hill and Wassenberg, 2000), and that nearly all prawn trawler discards are dead (Hill and Wassenberg, 1990).

Although there is a possibility of some species misidentification in the onboard samples, most of the fish species were recorded in both the on-shore and onboard samples. Furthermore, the relative contribution of individual fish species to the total discarded bycatch was similar for both data sets, with no species

exceeding a percent difference of 4%. This is an indication that the sampling procedure is being followed consistently, and onboard observers are correctly identifying the catch.

Sciaenids are largely distributed in the Tugela Bank trawl catches (three of the top five most commonly recorded species in both the on-shore and onboard samples) further supports the reports of common occurrence of this family on coastlines close to river mouths. Species of this family are widely distributed in shelf waters of the tropical and sub-tropical Indian, Pacific and Atlantic oceans (Lowe-McConnell, 1962; Longhurst and Pauly, 1987; Sasaki, 1996) and are important components of fisheries in many countries (Fennessy, 1994). These species are usually regarded as 'indicator' species of penaeid fishing grounds (Pauly and Neal, 1985). Investigation into the biology of the three most commonly recorded species, *O. ruber*, *J. fuscolineatus* and *J. dorsalis*, showed that the two latter species use the Tugela Bank as a nursery area, accounting for their abundance in bycatches in the area (Fennessy, 2000). The suitability of the Tugela Bank as a nursery area (Fennessy, 2000) is enhanced by the high turbidity of the area. The turbidity is due to the many rivers that flow into the sea, which in turn contributes to mud deposition on the east coast of South Africa (Flemming and Hay, 1988). In addition, the Tugela Bank is situated in the Natal Bight, which provides protection from the rough seas which characterise the east coast; these qualities make the Tugela Bank suitable as a nursery area for sciaenids (Fennessy, 2000).

6.3 Potential effects of trawling

Bottom trawling is a source of chronic and widespread disturbance on the seabed of shallow seas (Schratzberger *et al.*, 2002). Studies in the North Sea have shown that trawling has profound effects on the ecosystem because trawlers remove large portions of the biomass of target and bycatch species, and because trawl gears have direct impacts on the substratum and associated biota (Jennings *et al.*, 1999). The disturbance to the substratum is because trawls are in direct contact with the seabed, and because gear-induced water movements lead to resuspension of surface sediments (Jennings *et al.*, 1999). This physical disturbance of the sediment can lead to loss of biological organisation and can reduce species richness (Hall, 1995).

This study examined the temporal impacts of bottom trawling on the fish community on the Tugela Bank after a 13 year interval, by comparing the initial (1989-1992) survey of the bycatch composition by Fennessy (1994b) with the current composition (2003-2004). The species presence from both surveys was similar although the relative contributions by individual species differed. Several sciaenids featured prominently in the discarded bycatch, with three species present in the top five most common fish species in the 2003-2004 survey, and four species in the top five in the 1989-1992 study. The species that showed the most change in relative abundance was *S. ruconius*, which featured in the top five most commonly recorded species in 2003-2004, replacing *A. nibe* in the 1989-1992 study. The high overall

relative abundance of *S. ruconius* was partly due to high contributions from a few samples (three samples contributed over 50% of the total numbers of this species). *S. ruconius* is a slow-swimming fish which sometimes swims in dense, slow-moving shoals over the seabed (van der Elst, 1981), which accounts for the high relative abundance in some samples.

Of the 15 most commonly recorded species in the 1990s, 11 showed a decline and four showed an increase in relative abundance. Seven of the eight most commonly recorded species in the 1990s showed a decline in relative abundance of 26-91%, which suggests that trawling may have had a negative effect on the abundance of these species over time. Three of the five sciaenid species featuring in the 1990s and 2000s data sets showed a decline in relative abundance with *A. nibe* declining the most (59%). In contrast, *O. ruber* and *A. thorpei* increased in abundance by 32% and 189% respectively.

Also of significance are the observed changes in frequency of occurrence (Table 2.5). Of the species which were commonly recorded in 1990s, those which showed the greatest decline in frequency of occurrence were demersal species. *A. nibe*, for example, was a ubiquitous species in the 1989-1992 samples, but only occurs in 62% of trawls in the 2003-2004 data. *L. guentheri* and *D. longimanus* both occurred in more than 80% of trawls in the 1990s, but currently only feature in about 45% of trawls.

Considering the species which exhibit the greatest changes over the 13-year period, it is apparent that most of the species that showed large increases in relative abundance or frequency of occurrence are pelagic, and the species that showed large decreases are demersal. This is perhaps to be expected because demersal species occur close to the seabed, which makes them more likely than pelagic species to be vulnerable to the effects of bottom trawling over time. The growth rate of the species also has an effect on their vulnerability to harvesting, i.e. if the species is slow growing (growth rate $k < 0.3 \text{ yr}^{-1}$), there is a higher probability of decreased abundance over time as a result of trawling. This is because stocks of short-lived, early maturing species are less affected by trawling, whereas long-lived, late-maturing species are more vulnerable (Alverson *et al.*, 1994; Stobutzki *et al.*, 2001). Fecundity can also affect the species' vulnerability to trawling because if low numbers of larvae are produced, fast stock recovery would be difficult. Seasonality also affects vulnerability, since species which occur more commonly in the main trawling season (April-July) are more vulnerable than species which are more common out of the main trawling season. In order to relate the observed changes in abundance/occurrence to a species' vulnerability to trawling, a simple index of change was derived, by taking the product of the absolute change in relative abundance and the frequency of occurrence (Table 5). For the species which showed the most change in abundance and/or frequency of occurrence, Table 5 illustrates how the biology of these species contributes to their vulnerability to trawling and attempts to explain the increased relative abundance of pelagic species over time.

Table 5: Species that showed the most change in relative abundance and/or frequency of occurrence over the 13-year period, with characteristics which determine their resilience to trawling. The index of change is a product of absolute change in relative abundance and frequency of occurrence.

Scientific name	Change in relative abundance	Change in frequency of occurrence	Index of change	Pelagic (P) Demersal (D)	Shoaling	Slow Growing (K<0.3 yr ⁻¹)	Migration	Fecundity	Resilience to Trawling ¹
<i>Secutor ruconius</i> ¹	9	28	237	P	YES	NO	YES	HIGH	HIGH
<i>Ariobucca nibe</i> ²	-5	-31	-159	D	YES	YES	YES*	LOW	LOW
<i>Pomatomus saltatrix</i> ³	3	56	152	P	NO	YES	YES	HIGH	HIGH
<i>Drepane longimanus</i> ⁴	-3	-40	-122	D	NO	YES	YES	LOW	LOW
<i>Lagocephalus guentheri</i> ¹	-3	-38	-118	D	NO?	NO	NO?	HIGH	HIGH?
<i>Alepes djedaba</i> ⁵	2	45	89	P	YES	NO	NO	HIGH	HIGH
<i>Leiognathus equula</i> ⁵	2	32	75	P	YES	NO	NO	HIGH	HIGH
<i>Trichiurus lepturus</i> ^{5, 6}	-4	-15	-64	P/D	YES	YES?	YES?	LOW	LOW
<i>Thyssa vitirostris</i> ⁵	-4	-15	-62	P	YES?	NO	NO	HIGH	HIGH
<i>Pellona ditchela</i> ¹	2	17	32	P	YES	NO	NO	HIGH	HIGH

a. Juveniles occur on the Tugela Bank, sub-adults and adults move to deeper waters.

¹ Froese and Pauly (2005)

² Fennessy (2000)

³ Govender (1996)

⁴ Hoda and Khan (1994)

⁵ van der Elst (1981)

⁶ Kwok and Ni (1999)

The biological characteristics of these species i.e. whether they are pelagic or demersal, shoaling, slow-growing, migratory or highly fecund, were established from the literature or assigned a rating based on the most likely scenario. Theoretically, population dynamics suggests that slow growing species, which grow to large sizes and mature late in their life cycle, should decrease faster due to fishing pressure compared with fast growing, early maturing species (Southwood, 1976; Hoenig and Gruber, 1990). The species that showed a positive index of change i.e. they showed an increase in their relative abundance and/or frequency of occurrence, have high resilience to trawling. In contrast, species which had a negative index of change have low resilience to trawling, with the exception of *L. guentheri* and *T. vitirostris*. However there is some uncertainty about the resilience of *L. guentheri* as Fishbase (2005) assigns the species a theoretical high growth rate of 0.46 yr⁻¹ based on small maximum size (26cm) whilst Smith and Heemstra (1986) reports a maximum length of at least 41cm, which implies a slower growth rate. This species then, may in fact be less resilient than indicated in the table. The decline in *T. vitirostris* is surprising given it's high resilience, but this species is highly estuarine dependent (Whitfield, 1999) and is particularly abundant in the St Lucia Estuary (Wallace, 1975). This estuary mouth has been closed for 3 years, so it is possible that recruitment of this species to the Tugela Bank has been affected, which in turn would reduce abundance in trawl catches.

Table 5 re-emphasises the resilience of pelagic species to trawling and the vulnerability of demersal species. Notwithstanding the apparent decline of several species of demersal fish on the Tugela Bank as

a result of trawling, it is not necessarily true to say that the increase in some pelagic species is due to trawling. Pelagic species frequently demonstrate large shifts in abundance due to recruitment failures or success. This possibility cannot be ruled out for pelagic species such as *S. ruconius*, *P. saltatrix* and *L. equula* on the Tugela Bank.

Few other studies have examined the potential impacts of trawling over a long period. An Australian study also found that there was a decline in demersal species as a result of trawling because of their high susceptibility to capture by trawls (Stobutzki *et al.*, 2001). The species that are less susceptible to capture by trawls were generally pelagic with a broad depth distribution and wider range in the trawl grounds (Stobutzki *et al.*, 2001). These species also have a greater recovery capacity from the impacts of trawling. Harris and Poiner, (1991) compared the changes in abundance of fish species over a 20-year period of prawn trawling in the Southern Gulf of Carpentaria, Australia, and discovered that the total fish abundance had decreased. This study showed that although there was an overall decrease in total fish abundance, the abundance of most fish taxa had not changed. A total of 18 taxa had decreased, 12 had increased in relative abundance and 52 showed no significant change. The taxa that decreased were mostly benthic/demersal species, while those that increased were benthic-pelagic species, as in the current study. Tiews (1983) analysed the German shrimp fishery for 27 years between 1954-1981 to assess the stock sizes of 25 indigenous fish and crustacean species. This study showed a decline in the species abundance of seven species, while three species increased in abundance. Another study by Philippart (1996) showed a reduction of demersal fish and benthic invertebrate species to very low levels of abundance over 35 years, as a result of bycatch in the bottom trawl fishery of the southeastern North Sea.

6.4 Estimated bycatch quantities

Gear size and season had a significant effect on the bycatch discard rate and although gear size was coded as large, medium and small trawl sizes, it is a much more complex factor than the length of the footrope. The number of nets, number of seams, number of floats, trawl door size, trawl speed, etc. all influence the effective net opening, and therefore the potential catch rates. For example, although the chico 500 and tri-net are both coded as large nets, the chico 500 has a larger vertical opening than the tri-net. In the absence of controlled experimental conditions, it has not been possible to determine inter-gear calibration co-efficients, so a simple proxy (gear size) was used. Comparison of catch rates between gear sizes, seasons and depths should therefore be treated with caution, and the estimation of total discard quantities is therefore also only an approximation. Consequently, differences in discard rates between the various depths and seasons will not be discussed further.

Some species that are caught as bycatch on prawn trawlers are marketable and therefore retained. This component of the bycatch is referred to as retained bycatch, and includes fish like *A. thorpei*, *O. ruber*, *S. sagax*, *P. saltatrix* and *P. kakaan*. The estimated total discarded bycatch for 2003 was about 89 tonnes,

which is much lower in comparison to the bycatch quantities estimated in the 1989-1992 study, which was 357 tonnes per year. However, the fleet effort for 1989-1992 was between 6 601 and 12 457 hours per year, whereas in 2003 the fleet effort was 1 032 hours. The reduced effort is partially because of the closed season between November-February, but also because only three trawlers were operating during the course of the current study. There was about 19 tonnes of retained catch in 2003 with the target species *F. indicus* contributing 10 tonnes to the retained catch, i.e. the prawn/discarded bycatch ratio is 1:9, which is similar to that estimated in the 1989-1992 survey, which was between 1:6 and 1:9. Other studies that have shown similar prawn/discarded bycatch ratios are by Young and Romero (1979) in the Gulf of California (1:15); Perez Mellado *et al.* (1982) in the Gulf of California (1:9.8); Pender *et al.* (1993) in Northern Australia (1:8-1:21); Kennelly *et al.* (1998) in New South Wales (1:10.4); Brewer *et al.* (1998) in the Gulf of Carpentaria (1:15.7-1:19.4); Cascorbi (2004) in the Gulf of Mexico (1:5.3) and the United States South Atlantic Regions (1:4.5).

6.5 Length Frequencies

O. ruber, *J. fuscolineatus*, *J. dorsalis* and *A. nibe* caught in the bycatch during this study were relatively small, which is typical of prawn trawl bycatch (Saila, 1983). The codend mesh in the prawn trawl fisheries is generally between 35 and 55 mm in order to retain prawns, resulting in bycatch of various size classes (Fennessy, 1994b). The predominance of smaller length classes in the catch is because the Tugela Bank functions as a nursery area for these and other demersal species. The sizes of these sciaenid species at 50% maturity for both sexes are 237 mm (*O. ruber*), 125 mm (*J. fuscolineatus*) and 168 mm (*J. dorsalis*) respectively (Fennessy, 1994b; Fennessy, 1999). Most of the individuals of *O. ruber* and *J. dorsalis* caught in the 1990s and 2000s, were immature (Figure 2). The majority of *J. fuscolineatus* caught in the bycatch from the 1990s were immature, whilst most of the individuals caught in the 2000s are larger than the size at 50% maturity. Although the mean sizes of all three species has increased in the past 13 years, this is probably a result of a gear change, since mesh size increased from 35mm in the 1990s to 55mm in the 2000s. However *J. dorsalis* and *J. fuscolineatus* showed an increase in minimum size and a decrease in the maximum size caught. This suggests that there are fewer large fish of these species, which could be a sign of overexploitation.

7. MANAGEMENT OF THE FISHERY AND RECOMMENDATIONS

Globally bycatch in prawn trawl fisheries is problematic and effective management of these fisheries is necessary to ensure the sustainability of both the target and bycatch species. The high abundance and diversity of other organisms in areas where penaeid trawling occurs requires a multispecies management approach to reduce bycatch without reducing the prawn catch. Although this is a difficult task, there are various management measures that have attempted to reduce bycatch both in South Africa and internationally. The South African Marine Living Resources Act (1998) includes bycatch issues in its

objectives in that there is a ‘‘need to protect the ecosystem as a whole, including species that are not targeted for exploitation.’’ The United Nations’ Law of the Sea Convention (1994), to which South Africa is a signatory, recognised the need for the conservation, protection and co-ordinated management of the marine and coastal environment and its resources. Since bycatch is a growing issue, it is recommended that it be included in more detail in both South African and international legislation. This may include a requirement for fishing companies to make every effort to reduce bycatch, eg. the use of Bycatch Reducing Devices (BRDs). These devices improve selectivity, thus reducing bycatch quantity (Broadhurst, 2000) and many of the world’s prawn trawl fisheries are required to use BRDs (Tillman, 1993). BRDs protect the target species of the prawn trawl fisheries and its juveniles, and also the target species of other fisheries (Broadhurst and Kennelly, 1995). Imposing maximum limits on the fishery specific targets:bycatch ratios could also be included in the legislation. The implementation of legislated management measures is a difficult task and it is a challenge to enforce these laws.

From the results of this study, there is evidence that there are species that have shown an increase and others a decline in relative abundance over the 13 years since the initial survey. Stobutzki *et al.* (2001) noted that bycatch threatens the sustainability and recovery capacity of certain species groups, which results in shifts in species dominance. This study shows that there has been an increase in the relative abundance of pelagic species. This shift in species abundance in turn potentially threatens the populations of certain species, and threatens the sustainability of the ecosystem as a whole. Trawling effort is currently restricted to a maximum of five trawlers operating at any one time on the Tugela Bank and there is a closed season between November and February. However there are some precautions that could be taken in order to reduce the amount of bycatch caught on prawn trawlers. For example, the compulsory installation of BRDs on trawlers should be researched, tested and implemented as soon as possible, as this will go a long way in reducing the bycatch volumes. In some countries (Australia) BRDs are used in conjunction with other management measures such as seasonal, spatial and daytime closures, limited entry and gear restrictions designed to control fishing effort (Brewer *et al.*, 1998). Countries like the United States of America (USA) have gone as far as to ban imports of penaeids from countries that do not use or enforce the use of BRDs on their commercial prawn trawl fisheries (Brewer *et al.*, 1998). The compulsory use of BRDs has proved highly successful in Norway (Isaksen, 2000), Australia (Kennelly, 2000), Mexico (Garcia-Caudillo *et al.*, 2000). Studies on the effectiveness of BRDs on prawn trawlers in the USA have also proved that BRDs not only significantly lower fish biomass caught on prawn trawlers, but also catch more prawns (Rogers *et al.*, 1997).

RECOMMENDATIONS

- It is recommended that more discarded bycatch samples be taken in deep waters (>30m) and during the night, so that ongoing comparisons can be made with historic data (Fennessy, 1993).
- Although it was not demonstrated conclusively in this study, gear type appears to have an impact on the discard bycatch rate, therefore it is advisable to do more research on gear types in order to determine which gear catches prawns optimally and reduces bycatch quantities.
- Studies should be undertaken on the effectiveness of BRDs and BRD use should ultimately be made compulsory for all prawn trawlers operating on the Tugela Bank.
- In addition, to reduce bycatch, a closed season from the 1st September until the 31st October should be implemented, as at this time prawn catches are low.
- The observer programme should continue in order to assess if the changes in species composition. Ongoing education of the observers on species identification is required for improved data quality.
- There should be more comprehensive assessment studies on the most commonly caught bycatch species, to determine their stock status and to assess their ability to withstand fishing pressure.

REFERENCES

- Alverson, D.L.; Freeberg, M.H.; Murawski, S.A. and Pope, J.G. (1994) – A global assessment of fisheries bycatch and discards. *FAO Fish. Tech. Pap.* No. 339, 233 pp.
- Andrew, N.L. and Pepperell, J.G. (1992) – The bycatch of shrimp trawl fisheries. *Oceanogr. Mar. Biol. Ann. Rev.* **30**, pp. 527-565.
- Barnard, K.H. (1950) – Descriptive catalogue of the South African decapod crustacea. *Ann. S. Afr. Mus.* **38**, 837 pp.
- Begg, G.W. (1984) – The comparative ecology of Natal's smaller estuaries. *Natal Provincial Town and Regional Planning Report (Pietermaritzburg)* **62**, pp. 1-199.
- Blaber, S.J.M.; Brewer, D.T.; Salini, J.P. and Kerr, J. (1990) – Biomasses, catch rates and abundances of demersal fishes, particularly predators of prawns in a tropical bay in the Gulf of Carpentaria, Australia. *Mar. Biol.* **107**, pp. 397-408.
- Brewer, D.; Rawlinson, N.; Eayrs, S. and BurrIDGE, C. (1998) – An assessment of bycatch reduction devices in a tropical Australian prawn trawl fishery. *Fish. Res.* **36**, pp. 195-215.
- Broadhurst, M.K. and Kennelly, S.J. (1995) – A trouser-trawl experiment to assess codends that exclude juvenile mullet (*Argyrosomus hololepidotus*) in the Hawkesbury River prawn-trawl fishery. *Mar. Freshw. Res.* **46** (6), pp. 953-958.
- Broadhurst, M.K. (2000) – Modifications to reduce bycatch in prawn trawls: A review and framework for development. *Rev. Fish. Biol. Fish.* **10**, No. 1, pp. 27-60.
- Cascorbi, A. (2004) – Seafood Watch: Wild-Caught warmwater shrimp (the penaeid shrimps) in the Gulf of Mexico and U.S. South Atlantic Region. Seafood Report, Monterey Bay Aquarium, 52 pp.
- Colura, R.L. (2001) – Effect of the salt-box catch-bycatch separation procedure, as used by the Texas shrimp industry, on short-term survival of bycatch. *Fish. Bull.* **99** pp. 399-409.
- DEAT: Prawn Fishery (2002) – In *Department of Environmental Affairs and Tourism, South Africa*, (ed.). Where have all the fish gone: Measuring transformation in the South African fishing industry, pp. 28.

- Evans, S.M. and Wahju, R.I. (1996) – The shrimp fishery of the Arafura Sea (Eastern Indonesia). *Fish. Res.* **26** (3-4) pp.
- FAO (1995) – *Fisheries Statistics, Catches and Landings*, Vol. **80**, 713 pp.
- FAO (2002) – *Yearbook of fishery statistics*, Vol. **94**, FAO, Rome, 641 pp.
- Fennessy, S.T. (1993) – An investigation of the ichthyofaunal bycatch of the Tugela Bank prawn trawlers. MSc Thesis, University of Natal, Durban: 104 pp.
- Fennessy, S.T. (1994a) – Incidental capture of elasmobranchs by commercial prawn trawlers on the Tugela Bank, Natal, South Africa. *S. Afr. J. mar. Sci.* **14**, pp.287-296.
- Fennessy, S.T. (1994b) – The impact of commercial prawn trawlers on linefish off the north coast of Natal, South Africa. *S. Afr. J. mar. Sci.* **14**, pp. 263-279.
- Fennessy, S.T.; Villacastin, C. and Field, J.G. (1994) – Distribution and seasonality of ichthyofauna associated with commercial prawn trawl catches of the Tugela Bank of Natal, South Africa. *Fish. Res.* **20**(2/3), pp.263-282.
- Fennessy, S.T. (1995) – Relative abundances of non-commercial crustaceans in the bycatch of Tugela Bank prawn trawlers off KwaZulu-Natal, South Africa. *Lammergeyer* **43**, pp. 1-5.
- Fennessy, S.T. and Groeneveld, J.C. (1997) – A review of the offshore trawl fishery for crustaceans on the east coast of South Africa. *Fish. Man. and Ecol.* **4**, pp. 135-147.
- Fennessy, S.T. (2000) – Aspects of the biology of four species of Sciaenidae from the East Coast of South Africa. *Estua. Coas. & Shel. Sci.* **50** (2), pp. 259-269.
- Fennessy, S.T. (2002) – Preliminary investigation of a bycatch reduction device for the Tugela bank fishery in KwaZulu-Natal. ORI Unpublished Report 203, 10pp.
- Flemming, B. and Hay, R. (1988) – Sediment distribution and dynamics on the Natal continental shelf. In *Coastal ocean studies off Natal, South Africa*, Schumann, E.H. (ed.) Berlin, Springer-Verlag: pp. 47-80.
- Froese, R. and Pauly, D. (eds) (2005) – *Fishbase World Wide Web* electronic publication. www.fishbase.org. version (03/2005).

- Garcia-Caudillo, J.M., Cisneros-Mata, M.A. and Balmori-Ramirez, A. (2000) – Performance of a bycatch reduction device in the shrimp fishery of the Gulf of California, Mexico. *Biol. Conserv.* **92** (2) pp. 199-205.
- Govender, A. (1996) – Mark-Recapture models for determination of mortality, migration and growth in *Pomatomus saltatrix* (Teleostei). PhD thesis, University of Natal.
- Hall, M.A. (1995) – A classification of bycatch problems and some approaches to their solutions. *Fish. Cent. Res. Rep.* **2**, pp. 65-74.
- Harris, A.N. and Poiner, I.R. (1991) – Changes in species composition of demersal fish fauna of Southeast Gulf of Carpentaria, Australia, after 20 years of fishing. *Mar. Biol.* **111**, pp.503-519.
- Hill, B.J. and Wassenberg, T.J. (1990) – Fate of discards from prawn trawlers in Torres Strait. *Aust. J. Mar. Freshw. Res.* **41**, pp. 53-64.
- Hill, B.J. and Wassenberg, T.J. (2000) – The probable fate of discards from prawn trawlers fishing near coral reefs. A study in the northern Great Barrier Reef, Australia. *Fish. Res.* **48** pp. 277-286.
- Hoda, S.M.S. and Khan, M.I. (1994) – Some aspects of the reproductive biology of *Drepane longimana*. *Mar. Res.* **3** (2), pp. 47-55.
- Hoening, J.M. and Gruber, S.H. (1990) – Life history patterns in the elasmobranchs: Implications for fisheries management. In: *Advances in the Biology, Ecology, Systematics and the Status of the Fisheries* (eds H.L. Pratt, S.H. Gruber and T. Tanuchi). *NOAA Technical Report*, **90**, pp. 1-16.
- Hunter, I. (1988) – Climate and weather off Natal. In *Coastal ocean studies off Natal, South Africa*, Schumann, E.H. (ed.). New York, Springer Verlag. pp. 81-100.
- Isaksen, B. (2000) – The Norwegian procedure and experience concerning acceptance of new selective technologies and practices. *FAO Fisheries Report*, Rome, no. **588**, pp. 98-105.
- Japp, D.W. (1997) – Discarding practices and bycatches for fisheries in the southeast Atlantic region (area 47). *FAO Vol.* **547**, pp. 235-256.
- Jennings, S.; Alvsvag, J.; Cotter, A.J.R.; Ehrich, S.; Greenstreet, S.P.R.; Jarre-Teichmann, A.; Mergardt, N.; Rijnsdorp, A.D. and Smedstad, O. (1999). – Fishing effects in northeast Atlantic shelf seas:

patterns in fishing effort, diversity and community structure. III. International trawling effort in the North Sea: an analysis of spatial and temporal trends. *Fish. Res.* **40**, pp. 125-134.

Kennelly, S.J.; Kearney, R.E.; Liggins, G.W. and Broadhurst, M.K. (1993) – The effect of shrimp trawling bycatch on other commercial and recreational fisheries – An Australian perspective. In *Proceedings of the International Conference on Shrimp Bycatch, Lake Buena Vista, Florida, 1992*. Tallahassee, Florida, Southeastern Fisheries Association. pp. 97-113.

Kennelly, S.J.; Liggins, G.W. and Broadhurst, M.K. (1998) – Retained and discarded by-catch from oceanic prawn trawling in New South Wales, Australia.

Kennelly, S.J. (2000) – The Australian procedure and experience with the introduction and acceptance of new sustainable fishing technologies. FAO Fisheries Report. Rome no. **588**, suppl., pp. 106-122.

Kwok, K.Y. and Ni I.H. (1999) – Reproduction of cutlassfishes, *Trichiurus* spp. From the South China Sea. *Mar. Ecol. Pr. Ser.* **176**, pp. 39-47.

Longhurst, A.R. and Pauly, D. (1987) – *Ecology of Tropical Oceans*. Academic Press, Inc., San Diego, 407 pp.

Lowe-McConnell, R.H. (1962) – The fishes of the British Guiana continental shelf, with notes on their natural history. *J. Linn. Soc. Lon. (Zool.)*, **44**, pp. 669-700.

Lutjeharms, J.R.E. and Connell, A.D. (1989) – The Natal Pulse and inshore counter currents off the South African East Coast. *S. Afr. J. Sci.* **85** (7), pp. 533-534.

Lutjeharms, J.R.E.; Grundlingh, M.L. and Carter, R.A. (1989) – Topography induced upwelling in the Natal Bight. *S. Afr. J. Sci.* **85**, pp. 310-316.

Maharaj, V. and Recksiek, C. (1991) – The by-catch from the artisanal shrimp trawl fishery, Gulf of Paria, Trinidad. *Mar. Fish. Rev.* **53**, pp 9-15.

McCormick, S.; Cooper, J.A.G. and Mason, T.R. (1992) – Fluvial sediment yield to the Natal coast: a review. *S. Afr. J. Aqua. Sci.* **18**(1/2), pp. 74-88.

Mann, B.Q.; Scott, G.M.; Mann-Lang, J.B.; Brouwer, S.L.; Lamberth, S.J.; Sauer, W.H.H. and Erasmus, C. (1997) – An evaluation of participation in management of the South African spear fishery. *S. Afr. J. Mar. Sci.* **18** pp. 179-194.

- Nance, J.M. and Scott-Denton, E. (1997) – Bycatch in the Gulf of Mexico shrimp fishery. CSIRO: Collingwood (Australia), pp. 98-102. (Developing and sustaining world fisheries resources. The state of science and management, Hancock, D.A. (eds), Smith, D.C. (eds), Grant, A. (eds), Beumer, J.P. (eds).
- Pauly, D. and Neal, R. (1985) – Shrimp vs. fish in Southeast Asian fisheries: the biological, technological and social problems. In: Recursos Pesqueros Potenciales de Mexico: La Pesca Acompañante del Camaron. Progr. Univ. de Alimentos, Inst. Cienc. Del Mar y Limnol. UNAM, Mexico, pp. 487-510.
- Pender, P.J.; Willing, R.S. and Ramm, D.C. (1993) – Northern Prawn Fishery bycatch study: distribution, abundance, size and use of bycatch from the N.T. (Australian Northern Territory) mixed species fishery. *Fish. Rep.* – Northern Territory Department of Primary Industry and Fisheries (Australia), **26**.
- Perez Mellado, J.; Romero, J.M.; Young, R.H. and Findley, L.T. (1982) – Yields and composition of bycatch from the Gulf of California. *Fish by-catch...Bonus from the sea* [Report of a technical consultation]. Ottawa (Canada): IDRC pp. 55-57.
- Philippart, C.J.M. (1996) – Long term impact of bottom fisheries on several bycatch species of demersal and benthic invertebrates in the southeastern North Sea. ICES: Copenhagen (Denmark) 22pp. *Counc. Meet. Of the Int. Counc. For the exploitation of the Sea*, Reykjavik (Iceland).
- Poiner I.R. and Harris A.N.M. (1996) – Incidental capture, direct mortality and delayed mortality of sea turtles in Australia's Northern Prawn Fishery. *Mar. Biol.* **125**, pp. 813-825.
- Rogers, D.R.; Rogers, B.D.; de Silva, J.A.; Wright, V.L. and Watson, J.W. (1997) – Evaluation of shrimp trawls equipped with bycatch reduction devices in inshore waters of Louisiana. *Fish. Res.* **33**, pp. 55-72.
- Saila, S.B. (1983) – Importance and assessment of discards in commercial fisheries. *FAO Fish. Circ.* No. **765**, 62 pp.
- Sasaki, K. (1996) – Sciaenid fishes of the Indian Ocean (Teleostei, Perciformes). *Memoirs of the Faculty of Science Kochi University Series D: Biology*, **16/17**, pp. 83-95.
- Schratzberger, M.; Dinmore, T.A. and Jennings, S. (2002) – Impacts of trawling on the diversity, biomass and structure of meiofauna assemblages. *Mar. Biol.* **140** pp. 83-93.

- Schumann, E.H. (1982) – Inshore circulation of the Agulhas Current off Natal. *J. Mar. Res.* **40**(1), pp. 43-55.
- Smith, M.M. and Heemstra, P.C. (Eds.) (1986) – *Smith's Sea Fishes*. Johannesburg; Macmillan: 1046pp.
- Smith, A.R. (2000) – Effects of trawling and dredging on the bottom habitat and lost fishing gear on the environment. FAO Vol. **588**, pp. 40-47.
- South African Marine Resources Act (1998) – Chapter 1 **2** (e), *Government Gazette* No. **18930**, Vol. 395, pp. 14.
- Southwood, T.R.E. (1976) – Bionomic strategies and population parameters. In: *Theoretical Ecology: Principles and Applications* (ed. R.M. May), pp. 26-48. Blackwell Science, Oxford.
- Stratoudaski, Y.; Fryer, R.J.; Cook, R.M., Pierce, G.J. and Coull, K.A. (2001) – Fish bycatch and discarding in Nephrops trawlers in the Firth of Clyde (west of Scotland). *Aquat. Liv. Resour.* **14** (5), pp. 283-291.
- Stobutzki, I.; Miller, M. and Brewer, D. (2001) – Sustainability of fishery bycatch: a process for assessing highly diverse and numerous bycatch. *Environ. Conserv.* **28** (2) pp. 167-181.
- Stobutzki, I.C.; Miller, M.J.; Heales, D.S. and Brewer, D.T. (2002) – Sustainability of elasmobranchs caught as bycatch in a tropical prawn (shrimp) trawl fishery. *Fish. Bull.* **4**, pp. 800-812.
- Tiews, K. (1983) – Severe changes of fish and crustacean stocks in the German Wadden Sea during the period 1954-1982. *Informationen für die Fischwirtschaft. Hamburg* **30** (3) pp. 144-146.
- Tillman, M.F. (1993) – By-catch – the issue of the 90's. In: R.P. Jones (Ed.) *Int. Conf. on Shrimp Bycatch*, May 1992, Lake Buena Vista, Florida, Southeastern Fisheries Association, Tallahassee, FA, pp. 13-18.
- Van der Elst (1981) – Species descriptions. In *A Guide to The Common Fishes of Southern Africa*. Borchert (Ed.), Cape Town, South Africa, C. Struik Publishers, pp. 197.
- Van Der Elst, R.P. (1988) – Shelf ichthyofauna of Natal. In *Coastal ocean studies off Natal, South Africa*. Lecture notes on coastal and estuarine studies: **26**, Schumann, E.H. (Ed.). New York, Springer-Verlag: pp. 209-225.

- Vaughan, D.S.; Seagraves, R.J. and West, K. (1991) – An assessment of the status of weakfish stocks, 1982-1988. In *Atlantic States Marine Fisheries Commission, Special Report 21*, Washington, DC., pp. 1-75.
- Vendeville, P. (1990) – Tropical shrimp fisheries: Types of fishing gear used and their selectivity. *FAO Fish. Tech. Pap.* pp. 1-75.
- Wallace, J.H. (1975) – Estuarine fishes of the East Coast of South Africa. I Species composition and length frequency distribution in the estuarine and marine environments. II Seasonal abundance and migrations. *Invest. Rep. Oceanogr. Res. Inst.* **40** pp. 1-72.
- Walter, U. (1997) – Quantitative analysis of discards from brown shrimp trawlers in the coastal area of the East Frisian Islands. *Archive of fishery and marine research* **45** (1), pp. 61-76.
- Watson, R.A.; Dredge, M.L.C. and Mayer, D.G. (1990) – Spatial and seasonal variation in demersal trawl fauna associated with a prawn fishery on the central Great Barrier Reef, Australia. *Aust. J. Mar. Freshw. Res.* **41**, pp. 65-77.
- Watson, J.W. (1999) – Summary report on the status of bycatch reduction devices development. NOAA, MS Lab, P.O. Drawer 1207, Pascagoula, MS 39567.
- Whitfield, J.H. (1994) – An estuary association classification for fishes of Southern Africa. *Sout. Afr. J. Sci.* **90** (7) pp. 411-417.
- Ye, Y.; Alsaffar, A.H. and Mohammed, H.M.A. (1999) – Bycatch and discards of the Kuwait shrimp fishery. *Fish. Res.* **45**, pp. 9-19.
- Young, R.H. and Romero, J.M. (1979) – Variability in the yield and composition of by-catch recovered from the Gulf of California shrimping vessels. *Trop. Sci.* **21** (4), pp. 249-264.

APPENDIX

List of associated fauna recorded from Tugela Bank prawn catches in 2003-2004 including a teleost list from the 1989-1992 survey.

TELEOSTEI (1989-1992)

<i>Acanthocephala indica</i>	BANDFISH
<i>Aesopia cornuta</i>	UNICORN SOLE
<i>Alectis ciliaris</i>	THREADFIN MIRROFISH
<i>Alectis indicus</i>	INDIAN MIRROFISH
<i>Antennarius striatus</i>	STRIPED ANGLER
<i>Apogon kallopterus</i>	SPINYHEAD CARDINAL
<i>Apogon nigripinnis</i>	BULLSEYE CARDINAL
<i>Argyrosomus thorpei</i>	SQUARETAIL KOB
<i>Ariomma indica</i>	INDIAN DRIFTFISH
<i>Ariosoma scheelei</i>	TROPICAL CONGER
<i>Atrobucca nibe</i>	BLACKMOUTH CROAKER
<i>Bregmaceros atlanticus</i>	CODLET
<i>Callionymus marleyi</i>	SAND DRAGONET
<i>Carangoides hedlandensis</i>	BUMPNOSE KINGFISH
<i>Carangoides malabaricus</i>	MALABAR KINGFISH
<i>Chelonodon pleurospilus</i>	BLAASOP BEAUTY
<i>Choridactylus natalensis</i>	THREE STICK STINGFISH
<i>Cociella heemstrai</i>	SPOTFIN FLATHEAD
<i>Cubiceps baxteri</i>	BLACK FATHEAD
<i>Cylichthys orbicularis</i>	BIRDBEAK BURRFISH
<i>Cynoglossus attenuatus</i>	FOUR-LINED TONGUEFISH
<i>Cynoglossus lida</i>	ROUGHSCALE TONGUEFISH
<i>Dactyloptena orientalis</i>	HELMET GURNARD
<i>Drepane longimanus</i>	CONCERTINA FISH
<i>Dysomma anguillare</i>	ARROWTOOTH EEL
<i>Epinephelus andersoni</i>	CATFACE ROCKCOD
<i>Galeichthys</i> sp.	CATFISH
<i>Gazza minuta</i>	TOOTHED SOAPY
<i>Gerres filamentosus</i>	THREADFIN PURSEMOUTH
<i>Halieutea spicata</i>	SPINY SEABAT
<i>Hilsa kelee</i>	RAZORBELLY
<i>Johnius dorsalis</i>	SMALL KOB
<i>Johnius fuscolineatus</i>	BELLFISH
<i>Lagocephalus guentheri</i>	BLACKBACK PUFFER
<i>Lagocephalus inermis</i>	SMOOTH PUFFER
<i>Leiognathus equula</i>	SLIMEY
<i>Lutjanus russellii</i>	RUSSEL'S SNAPPER
<i>Megalaspis cordyla</i>	TORPEDO FISH
<i>Mene maculata</i>	PONY FISH
<i>Minous coccineus</i>	ONE STICK STINGFISH

<i>Muraenesox bagio</i>	PIKE CONGER
<i>Otolithes ruber</i>	SNAPPER KOB
<i>Parachaeturichthys polynema</i>	TAIL-EYED GOBY
<i>Paraplagusia bilineata</i>	FRINGELIP TONGUEFISH
<i>Parastromateus niger</i>	BLACK POMFRET
<i>Pellona ditchela</i>	INDIAN PELLONA
<i>Polydactylus sextarius</i>	BUSTARD MULLET
<i>Pomadasys kaakan</i>	JAVELIN GRUNTER
<i>Pomadasys maculatum</i>	SADDLE GRUNTER
<i>Pomadasys olivaceum</i>	OLIVE GRUNTER
<i>Pomatomus saltatrix</i>	SHAD
<i>Pseudorhombus arsius</i>	LARGETOOTH FLOUNDER
<i>Pseudorhombus natalensis</i>	SMALLTOOTH FLOUNDER
<i>Sardinella albella</i>	GOLDSTRIPE SARDINELLE
<i>Sardinops sagax</i>	PILCHARD
<i>Saurida undosquamis</i>	LARGESCALE LIZARDFISH
<i>Scomber japonicus</i>	MACKEREL
<i>Scomberomorus plurilineatus</i>	NATAL SNOEK
<i>Secutor ruconius</i>	PUGNOSE SOAPY
<i>Serranus cabrilla</i>	COMBER
<i>Solea bleekeri</i>	BLACKHAND SOLE
<i>Stephanolepis auratus</i>	PORKY
<i>Terapon jarbua</i>	THORNFISH
<i>Terapon theraps</i>	STRAIGHT-LINED THORNFISH
<i>Thryssa setirostris</i>	LONGJAW GLASSNOSE
<i>Thryssa vitrirostris</i>	BONY
<i>Torquigener balteatus</i>	SLENDER PUFFER
<i>Trachurus delagoa</i>	AFRICAN MAASBANKER
<i>Trichiurus lepturus</i>	RIBBON FISH
<i>Trypauchen microcephalus</i>	COMB GOBY
<i>Upeneus vittatus</i>	YELLOW-BANDED GOATFISH
<i>Uranoscopus archionema</i>	STARGAZER
<i>Uroconger lepturus</i>	LONGTAIL CONGER

TELEOSTEI (2003-2004)

<i>Alepes djedaba</i>	SHRIMP SCAD
<i>Antennarius hispidus</i>	SHAGGY ANGLER
<i>Apogon quadrifasciatus</i>	TWOSTRIPE CARDINAL
<i>Archamia mozambiquensis</i>	MOZAMBIQUE CARDINAL
<i>Argyrosomus thorpei</i>	SQUARETAIL KOB
<i>Ariomma indica</i>	INDIAN DRIFTFISH
<i>Atrobucca nibe</i>	BLACKMOUTH CROAKER
<i>Bregmaceros atlanticus</i>	CODLET
<i>Callionymus marleyi</i>	SAND DRAGONET
<i>Carangoides dinema</i>	SHADOW KINGFISH
<i>Carangoides ferdau</i>	FERDY KINGFISH
<i>Carangoides malabaricus</i>	MALABAR KINGFISH
<i>Cheimerius nufar</i>	SANTER
<i>Cociella heemstrai</i>	SPOTFIN FLATHEAD
<i>Cyclichthys orbicularis</i>	BIRDBEAK BURRFISH
<i>Cynoglossus attenuatus</i>	FOUR-LINED TONGUEFISH
<i>Cynoglossus lida</i>	ROUGHSCALE
<i>Drepane longimanus</i>	CONCERINA FISH
<i>Dysomma anguillare</i>	ARROWTOOTH EEL
<i>Epinephelus andersoni</i>	CATFACE ROCKCOD
<i>Epinephelus rivulatus</i>	HALFMOON ROCKCOD
<i>Galeichthys sp.</i>	CATFISH
<i>Gazza minuta</i>	TOOTHED SOAPY
<i>Johnius dorsalis</i>	SMALL KOB
<i>Johnius fuscolineatus</i>	BELLFISH
<i>Lagocephalus guentheri</i>	BLACKBACK PUFFER
<i>Lagocephalus inermis</i>	SMOOTH PUFFER
<i>Lagocephalus scleratus</i>	SILVER STRIPED PUFFER
<i>Leiognathus equula</i>	SLIMEY
<i>Macroramphosus scolopax</i>	SLENDER SNIPEFISH
<i>Megalaspis cordyla</i>	TORPEDO SCAD
<i>Monocentris japonicus</i>	PINEAPPLE FISH
<i>Otolithes ruber</i>	SNAPPER KOB
<i>Pagellus bellottii natalensis</i>	SAND SOLDIER
<i>Parachaeturichthys polynema</i>	TAIL-EYED GOBY
<i>Paralichthodes algoensis</i>	MEASLES FLOUNDER
<i>Parapriacanthus ransonneti</i>	SLENDER SWEEPER
<i>Parastromateus niger</i>	BLACK POMFRET
<i>Pelates quadrilineatus</i>	TRUMPETER
<i>Pellona ditchela</i>	INDIAN PELLONA
<i>Polydactylus sextarius</i>	BUSTARD MULLET
<i>Pomadasyys kaakan</i>	JAVELIN GRUNTER
<i>Pomadasyys maculatum</i>	SADDLE GRUNTER
<i>Pomadasyys olivaceum</i>	OLIVE GRUNTER
<i>Pomatomus saltatrix</i>	SHAD
<i>Psettodes erumei</i>	ADALAH

<i>Pseudorhombus natalensis</i>	SMALLTOOTH FLOUNDER
<i>Pterois russelli</i>	PLAINTAIL FIREFISH
<i>Rastrelliger kanagurta</i>	SUGAR MACKEREL
<i>Rhabdosargus holubi</i>	CAPE STUMPNOSE
<i>Sardinops sagax</i>	PILCHARD
<i>Saurida undosquamis</i>	LARGESCALE LIZARDFISH
<i>Secutor ruconius</i>	PUGNOSE SOAPY
<i>Sillago sihama</i>	SILVER SILLAGO
<i>Solea bleekeri</i>	BLACKHAND SOLE
<i>Sphyræna acutipinnis</i>	SHARP-FIN BARRACUDA
<i>Sphyræna flavicauda</i>	YELLOW-FIN BARRACUDA
<i>Stephanolepis auratus</i>	PORKY
<i>Stolephorus punctifer</i>	BUCCANEER ANCHOVY
<i>Terapon jARBua</i>	THORNFISH
<i>Thryssa setirostris</i>	LONGJAW GLASSNOSE
<i>Thryssa vitrirostris</i>	BONY
<i>Trachinocephalus myops</i>	PAINTED LIZARDFISH
<i>Trichiurus lepturus</i>	RIBBON FISH
<i>Upeneus vittatus</i>	YELLOW-BANDED GOATFISH

ELASMOBRANCHII (2003-2004)

<i>Aetobatus narinari</i>	SPOTTED EAGLERAY
<i>Carcharhinus obscurus</i>	DUSKY SHARK
<i>Dasyatis chrysonata</i>	STINGRAY
<i>Dasyatis thetidis</i>	THORNTAIL STINGRAY
<i>Gymnura natalensis</i>	BUTTERFLY RAY
<i>Himantura draco</i>	DRAGON STINGRAY
<i>Himantura gerrardi</i>	SHARPNOSE STINGRAY
<i>Himantura uarnak</i>	MARBLED STINGRAY
<i>Holohalaelurus punctatus</i>	SPOTTED CATSHARK
<i>Mustelus mosis</i>	HARDNOSED SMOOTH-HOUND SHARK
<i>Pteromylaeus bovinus</i>	DUCKBILL RAY
<i>Rhinobatos annulatus</i>	LESSER GUITARFISH
<i>Rhinobatos holcorhynchus</i>	SLENDER GUITARFISH
<i>Rhinobatos leucospilus</i>	GREYSPOT GUITARFISH
<i>Rhinoptera javanica</i>	FLAPNOSE RAY
<i>Rhizoprionodon acutus</i>	MILK SHARK
<i>Rhynchobatus djiddensis</i>	GIANT GUITARFISH
<i>Scylliogaleus queketti</i>	FLAPNOSE HOUND SHARK
<i>Sphyrna lewini</i>	SCALLOPED HAMMERHEAD
<i>Squatina africana</i>	ANGEL SHARK
<i>Taeniura melanospilos</i>	ROUND RIBBONTAILRAY
<i>Torpedo sinuspersici</i>	MARBLED ELECTRIC RAY

CRUSTACIA: REPTANTIA (2003-2004)

Charybdis cruciata
Charybdis merguensis
Charybdis natator
Conchoecetes artificiosus
Doclea muricata
Dorippe lanata
Gonioneptunus africanus
Ixoides cornutus
Matt sp.
Platylambrus quemvis
Portunus hastatoides
Portunus sanguinolentus

CRUSTACEA: NATANTIA

Fenneropenaeus indicus
Panulirus homarus
Parapenaeopsis acclivirostris
Trachypenaeus curvirostris

STOMATOPODA

Lysiosquilla maculata
Squilla armata

CEPHALOPODA

Sepia vermiculata
Sepiella cyanea
Sepiella cyanea
Uroteuthis duvauceli
Uroteuthis duvauceli

GASTROPODA

Bufonaria crumena
Bufonaria crumena
Ficus ficus
Polynices didyma
Tonna allium

PELYCOPODA

Anadata natalensis
Eudolium pyriforme
Macra aequisulcata
 Unknown heart shaped clam